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# Risk assessment of pollutants in different environmental matrices.

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- Toxicology: basic principles;
- Regulatory toxicology: toxicological evaluation;
- Risk assessment procedure;
- New challenges in Risk assessment;
- Case studies: from dietary exposure; from the environment.

# Toxicology

- Before new industrial and consumer substances/products can be developed and marketed, toxicology is needed to make sure that they can be used safely;
- For substances already produced, used and present in the environment, toxicology helps to determine whether or not a risk is present and how to mitigate that risk for the population and the environment;



Evaluation of possible adverse effects on humans and possible risk for animals and the environment due to exposure to the substances/products in the workplace, home or environment



# Toxicology

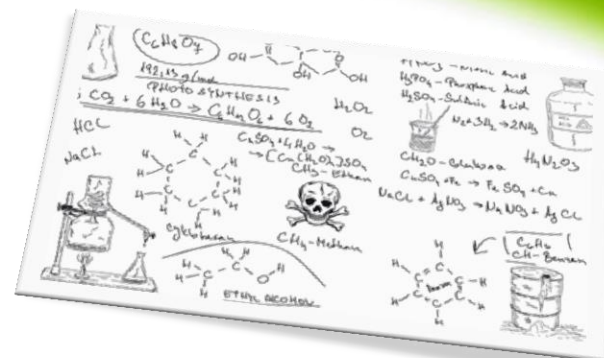
..anyone can become a toxicologist in two easy lessons, each of which takes ten years...(A.J.Lehman)



Interdisciplinary science to study the adverse effects of chemicals on living systems, including:

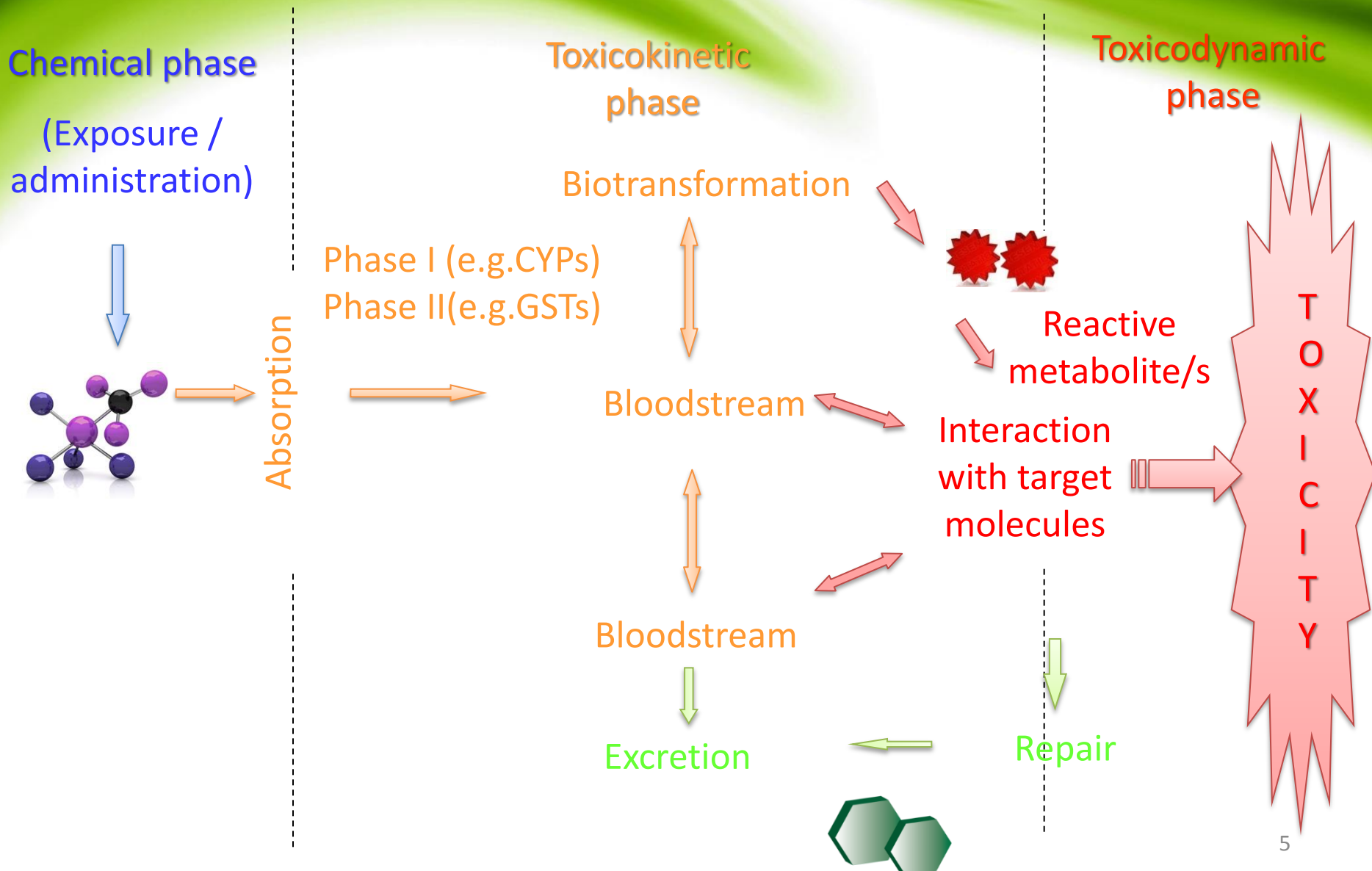
- Mechanisms of action and exposure to chemicals;
- Recognition, identification, quantification of hazards from occupational exposure to chemicals;
- Development of standards/regulations to protect humans and the environment from adverse effects of chemicals.

From phenomena observation and characterization... . testing  
(analysis , understanding, deduction )...  
to the quantification of the events...





# From the exposure to the target site



# *“The dose makes the poison”*

The basic principle:

*“All substances are poisons...the right dose differentiates a poison or a remedy”* → it is the dose that makes the poison

- Challenge for the toxicologist: to identify the dose(s) at which a specific substance can cause adverse effects;
- Differences between “toxic” or not: all substances are toxic, it is only the degree and type of toxicity different among all the agents.

Just because a chemical is present, does not mean there is a risk  
in the amount present

# The dose concept

By definition:

is the amount of a substance administered at one time.

Other parameters are needed to characterize the exposure to xenobiotics:

- the number of doses;
- frequency;
- total time period of the treatment.

i. A better means to allow for comparison of **effectiveness /toxicity** is the amount of a substance administered on a **body weight basis**



**mg/kg** = mg of substance per kg of bw

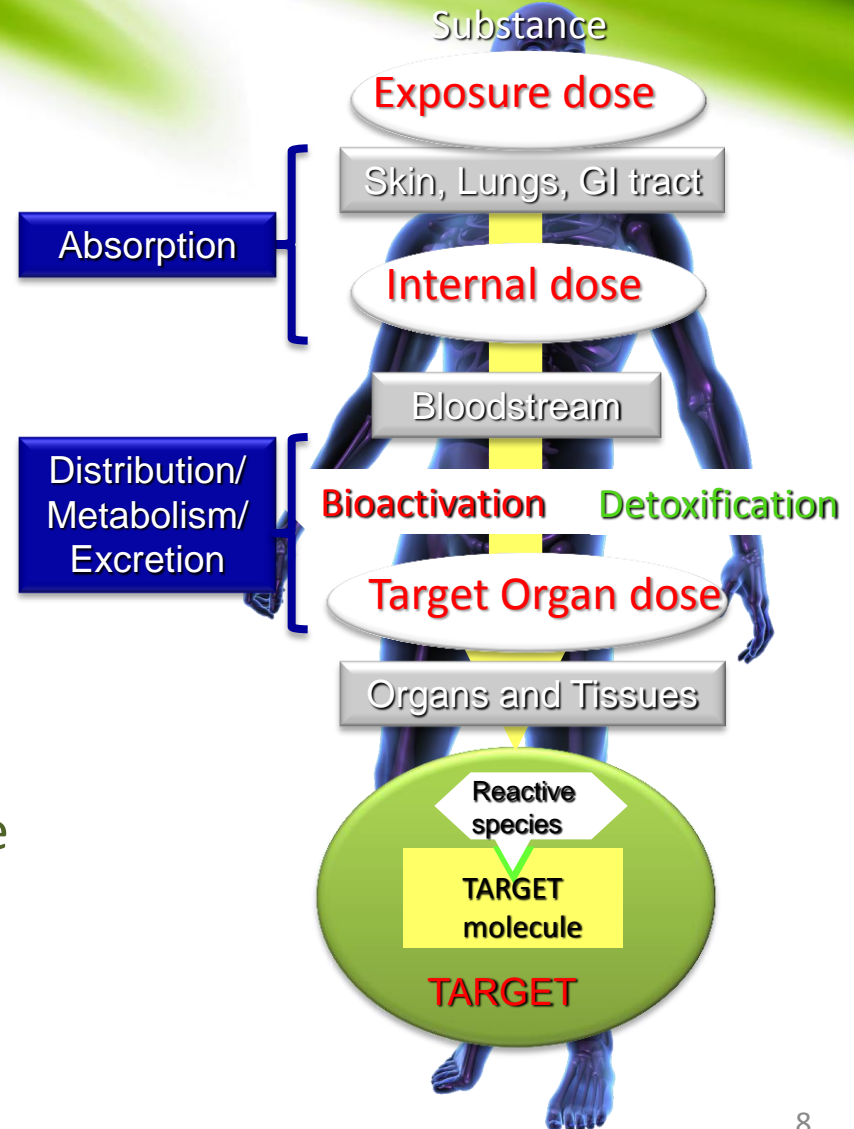
ii. the **time** over which the dose is administered. Important for several day or chronic exposures. Commonly used time unit = one day .



Usual dosage unit = **mg/kg/day**

# Relevant toxicologically doses

- The amount of a substance to which an individual/population is exposed (occupational/environmental):  
**Exposure dose;**
- The amount absorbed into the organism. The bioavailable dose that can cause the effect far from the exposure site:  
**Internal dose;**
- The amount able to reach the target site(s) and to cause the effect (adverse or effective):  
**Target organ/  
biologically effective dose**

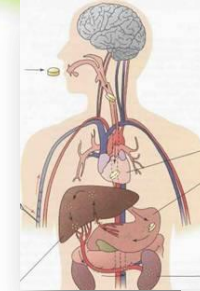




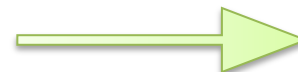
# Exposure/absorption measurement

Quantification: **exposure vs absorption**

- Various routes of exposure /absorption (e.g. oral, dermal, inhalation);



- Level of uncertainty in extrapolation the effects of absorbed dose from animals to humans



# The dose and its variable

Many variables are connected with the dose and consequently with the effects that occur at those doses:

- **Amount:** the magnitude of the dose;
- **Frequency:** how often (daily, weekly)
- **Duration:** how long? (acute, sub-chronic, chronic)
- **Route of exposure:** the way to be exposed (oral, dermal)
- **Individual variability/susceptibility:**  
subject characteristics (age, sex); health conditions (asthma)

The occurrence of the possible adverse effects due to the exposure to a chemical is evidenced during different toxicological studies carried out, taking into account all these variables.



Risk identification on humans,  
animals and environment

# Regulatory toxicology

**Regulatory toxicologists** evaluate the toxicity tests that is done on chemicals, contaminants and consumer products or preexisting information.

They must ask questions such as:

- ? What responses are considered “adverse”?
- ? To what doses are the consumers most likely to be exposed?
- ? They must define the risk associated with each chemical, and the level of risk that the public will accept.



The analysis of the nature and magnitude of risk is called  
**RISK ASSESSMENT**

# Toxicological studies

Accurate predictions of effects of chemicals on humans depend upon scientific studies.

Most toxicological studies are empirical in nature, and are performed on experimental animals (*in vivo*) or *in vitro* test systems (i.e., cell culture or other systems to mimic the results in part of an organism) or on the basis of *in silico* predictions.



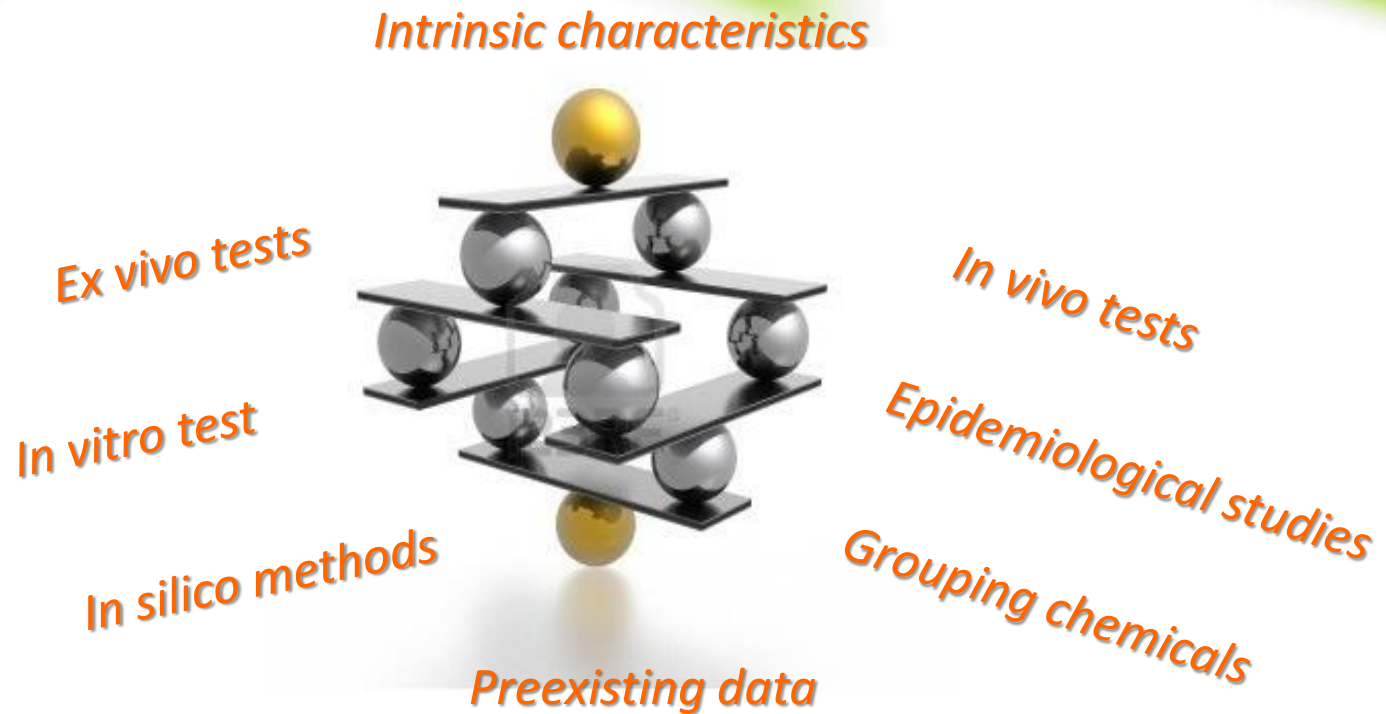
Since the results are often used for regulatory purposes, the goal of such studies is to predict effects in humans.



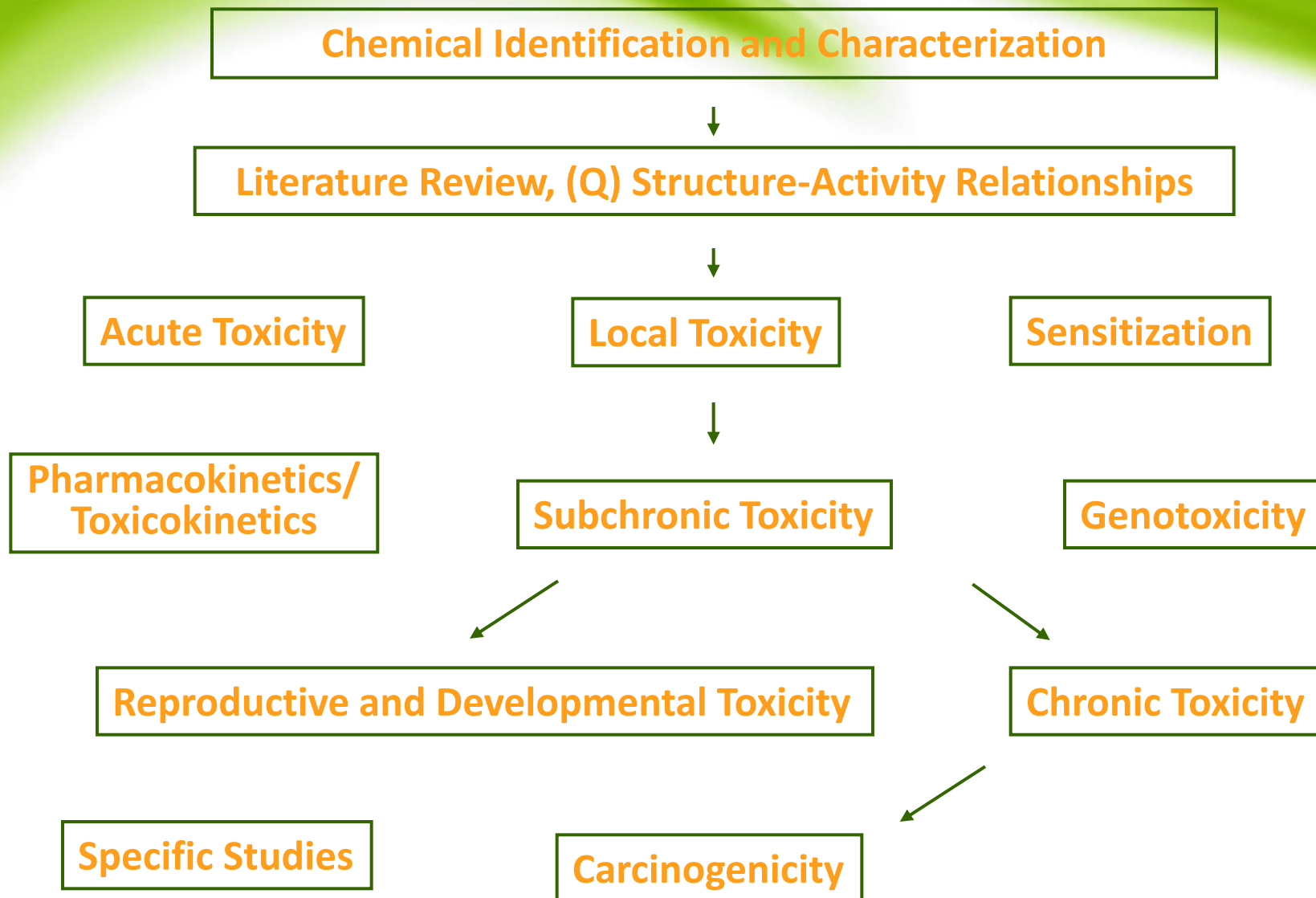
To achieve this goal, scientists need to understand the differences between experimental (animals or *in vitro* systems) and real conditions (in humans) in the way that they process xenobiotics, as well as the applicability to humans of results obtained.



# Toxicological evaluation

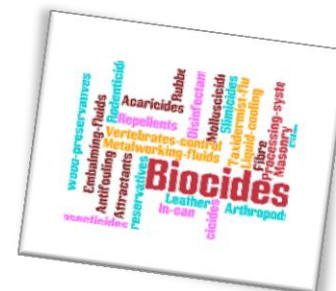
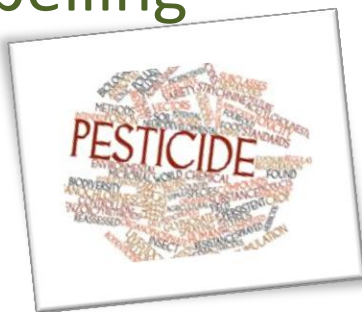


# A Typical Safety Evaluation Program



# INFORMATION REQUIREMENTS: ACTIVE SUBSTANCE

1. Identity of active substance
2. Physical and chemical properties
3. Further information (use, function, harmful organisms controlled, MoA)
4. Analytical Methods of detection and identification
5. **Toxicological and Metabolism studies**
6. (Residues in or on treated products, food and feed)
7. Fate and behaviour in the environment
8. Ecotoxicological studies
9. Literature data
10. Classification and labelling



## 5. TOXICOLOGICAL ENDPOINTS – core for a.s.

**5.1 Studies on absorption, distribution, metabolism and excretion** in mammals by oral (5.1.1) and other routes (5.1.2);

### **5.2 Acute toxicity:**

oral (5.2.1), dermal (5.2.2), inhalation (5.2.3);  
skin (5.2.4) and eye (5.2.5) irritation;  
skin sensitisation (5.2.6); phototoxicity (5.2.7);

Important for C&L (Classification and Labelling)

### **5.3 Short-term toxicity:**

oral 28-day study(5.3.1), oral 90-day study(5.3.2), other routes (5.3.3);

Identification of a NOAEL (No Observed Adverse Effect Level) and a LOAEL ( Low Observed Adverse Effect Level).

Preliminary to long term repeated toxicity studies and relevant for Reference values definition.

### **5.4 Genotoxicity testing:**

In vitro studies (5.4.1), in vivo studies in somatic cells (5.4.2), in vivo studies in germ cells (5.4.3);



## 5. TOXICOLOGICAL ENDPOINTS – core for a.s.

### 5.5 Long term toxicity and carcinogenicity

Identification of a NOAEL and LOAEL and major effects and target organ. Relevant for Reference values definition.

### 5.6 Reproductive toxicity:

Generational studies(5.6.1), Developmental toxicity studies (5.6.2);

Identification of a NOAEL and a LOAEL and major effects. Relevant for Reference values definition.

### 5.7 Neurotoxicity studies:

Neurotoxicity studies in rodents(5.7.1), Delayed polyneuropathy studies (5.7.2);

### 5.8 Other toxicological studies:

Toxicity studies of metabolites (5.8.1), Supplementary studies on the active substance (5.8.2), Endocrine disrupting properties (5.8.3); if necessary → mechanistic studies

### 5.9. Medical data



# RISK ASSESSMENT

# Risk assessment process

“A process intended to calculate or estimate the risk to a given target organism, system or (sub)population, including the identification of attendant uncertainties, following exposure to a particular agent, taking into account the inherent characteristics of the agent of concern as well as the characteristics of the specific target system”  
(joint OECD/IPCS survey)

## Risk Assessment:

Comparison of exposure with effects

Final outcome: Risk characterization



# What is risk assessment?

- Definition of **Hazard** : Inherent property of an agent or situation having the potential to cause adverse effects when an organism, system or (sub) population is exposed to that agent.
- Definition of **Exposure** : Concentration or amount of a particular agent that reaches a target organism, system or (sub) population in a specific frequency for a defined duration.
- Definition of **Risk**: The probability of an adverse effect in an organism, system or (sub) population caused under specified circumstances by exposure to an agent.

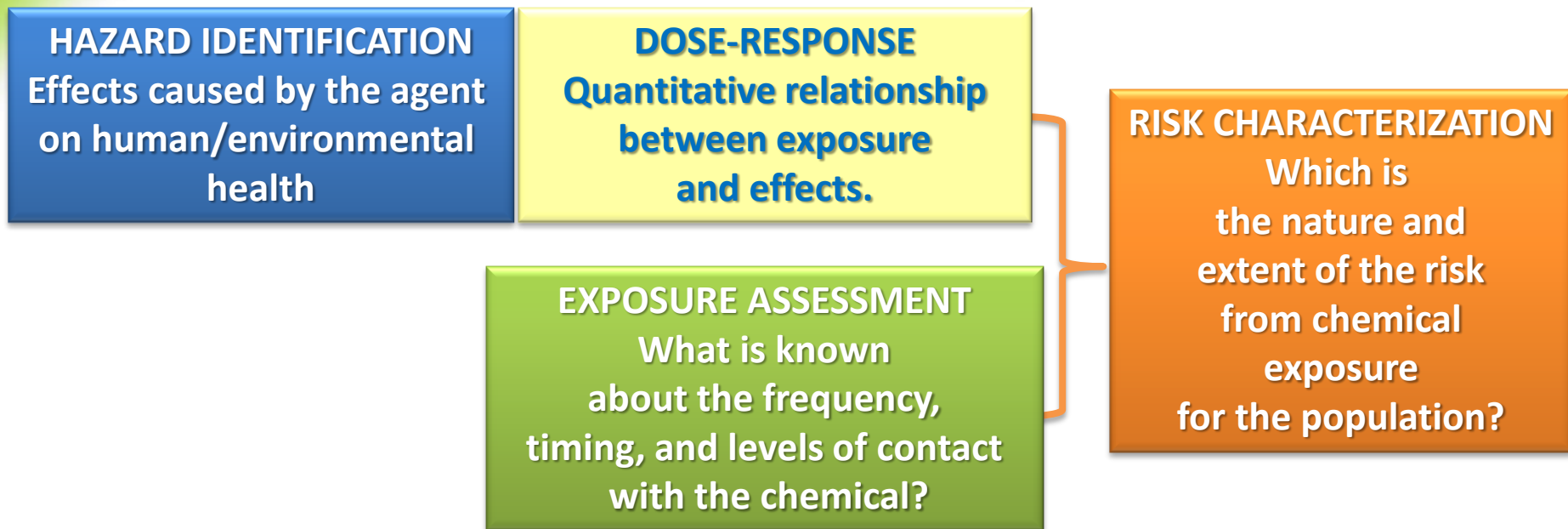


$$R = H_{\text{(HAZARD)}} \times E_{\text{(EXPOSURE)}}$$



# Toxicological Risk Assessment

A 4 Phases Process starting from data/info acquisition:



The procedure is defined at EU level: Commission Directive 93/67/EEC; Council Regulation (EEC) 793/93; TGD Technical Guidance Document on Risk Assessment, 2003; Commission Regulation (EC) No. 1488/94, and used by WHO, EPA, FDA, EFSA...



# HAZARD IDENTIFICATION

# I. Hazard identification

The hazard identification reflects the qualitative aspects of the assessment and provides answers to the question:



Which kind of adverse effects does the toxic agent induce?

**Hazard:** is an intrinsic feature of the toxic agent.

**Risk:** corresponds to the probability for a population of experiencing adverse effects once exposed to the toxic agent.



# How can hazard be identified?

- Acute toxicity studies on rats  
(single administration, oral & dermal LD<sub>50</sub>, inhalation LC<sub>50</sub>)
- Dermal and Eye Irritation (usually on rabbits)
- Sensibilization studies on guinea pig or LLNA on mice
- Mutagenicity test
- Repeated toxicity studies  
(ex: R48 Danger of serious damage to health by prolonged exposure;  
H373: May cause damage to organs through prolonged or repeated exposure)
- Carcinogenicity
- Reproductive/development toxicity studies



Classification and labeling for these endpoints must be performed according to regulation: **On hazard basis and not risk**

e.g. CLP Regulation (EU) n. 1272/2008

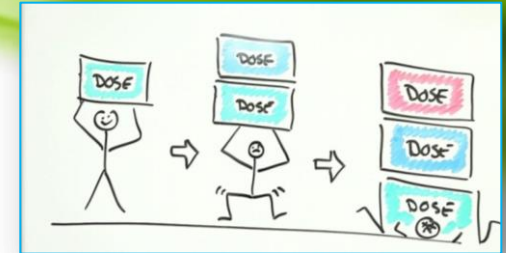


# DOSE-RESPONSE



## II. Dose-response Relationship

### Quantitative Aspect



At which concentration/ level does the adverse effect occur?

The threshold for the effect can be achieved with :

**Single exposure:** acute and generally at high doses  
→ accidents, poisonings.

**Repeated exposure:** low doses for prolonged times; doses generally non toxic if taken singularly  
→ cumulative effects

## Toxicological relevant compound → threshold dose

- Several adverse effects are caused by a substance only in the case the toxicological relevant compound (parent or toxic metabolite/s) reaches the threshold concentration at the target site/organ.
- Such a concentration is correlated to the exposure level (external) but mainly to the actual exposure (internal) of the organism (humans or animals).
- Therefore, the threshold concentration changes in relation to →
  - i. ≠ exposure route,
  - ii. ≠ speciesdue to ≠ toxicokinetics, ≠ mechanism of action

# TOXICOKINETIC

## Absorption, Distribution, Metabolism and/or Excretion

The **main routes of entries**:

GI mucosa (oral); Lung epithelium (inhalation); Skin (dermal)

The **main routes of excretion**:

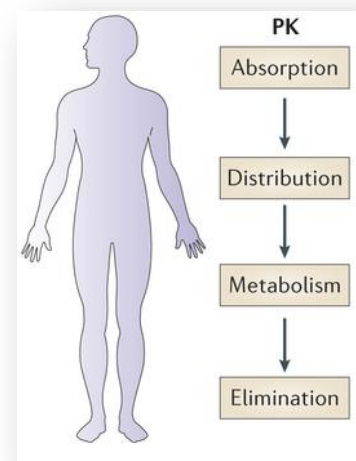
urinary ; faecal (via the bile); expired air; tears; sweat; breast milk

INTERNAL



At what concentration the chemical is present in different organ/tissue and in the target site?

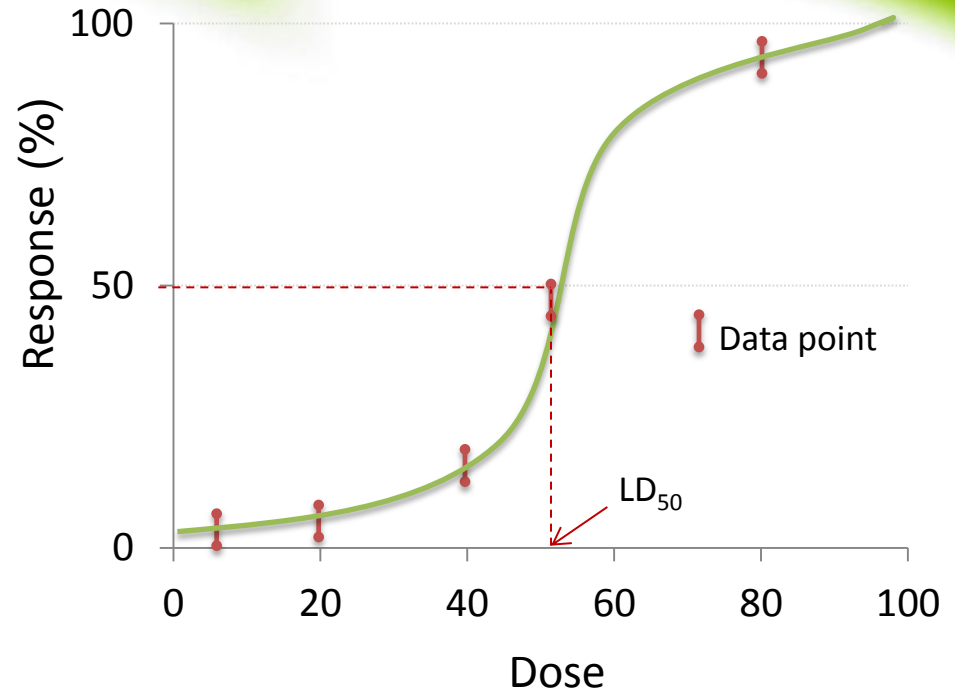
- Which the time-dependent fate of the chemical within the body?
- Which is the concentration of the tox relevant chemical species?



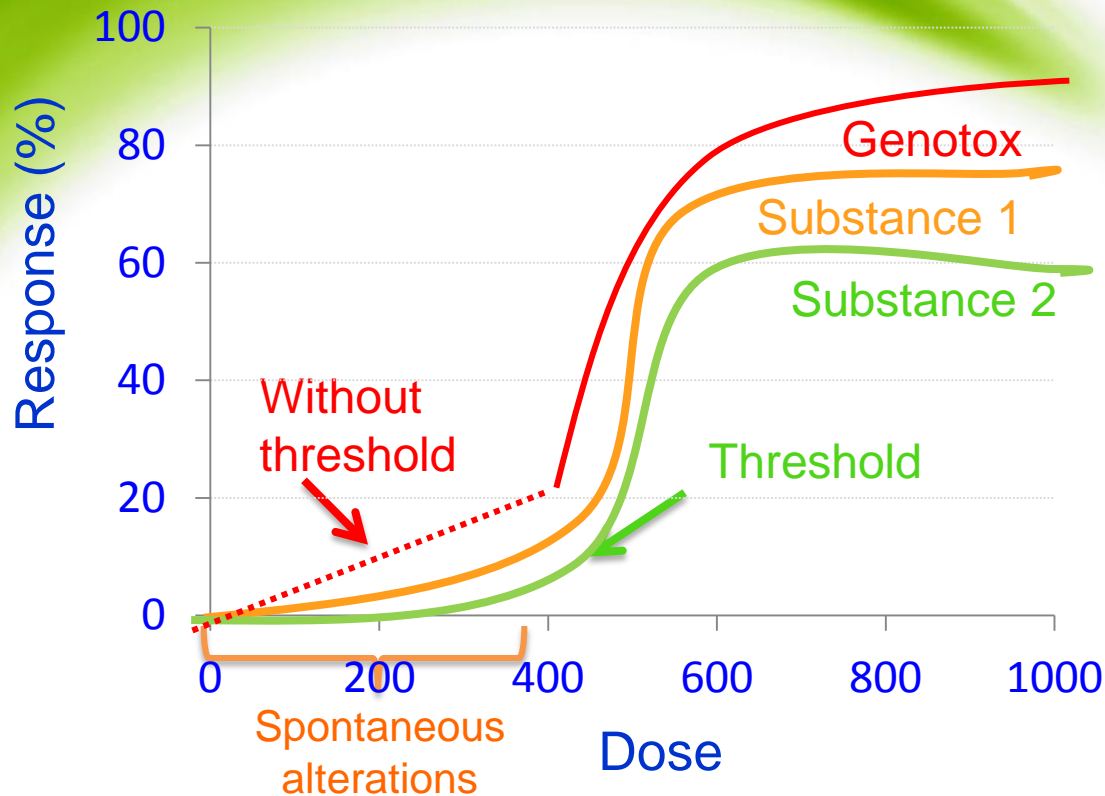
# Dose-response graph

➔ The data coming from the toxicological studies are plotted on a graph.

➔ The dose is plotted against the number or portion of animals exhibiting a specific response



# Responses to different agents vs increasing dose



The extent and nature of the effects related to:

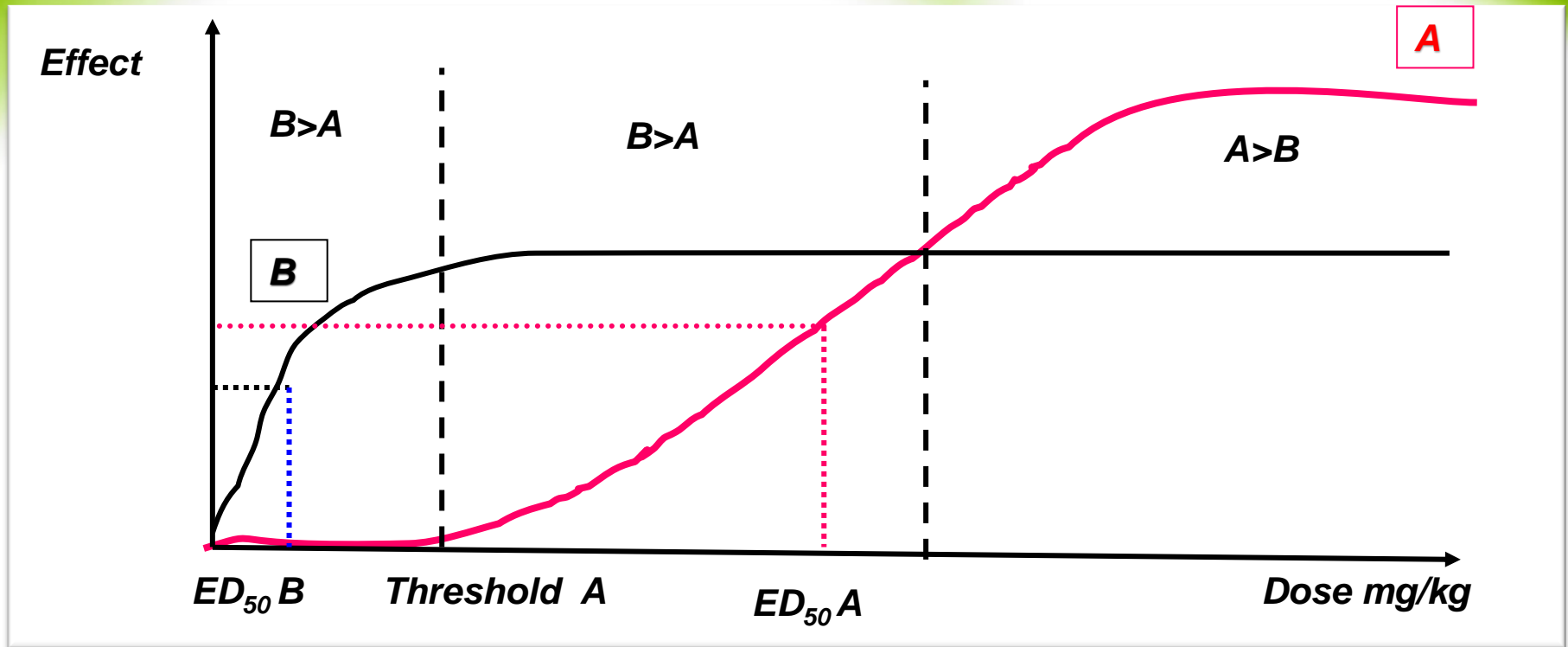
- dose;
- route of exposure;
- characteristics of the individual;
- time course and duration of the administration/exposure;
- spacing between doses

Critical variable in determining whether/how or not adverse effects occur.



# Threshold dose

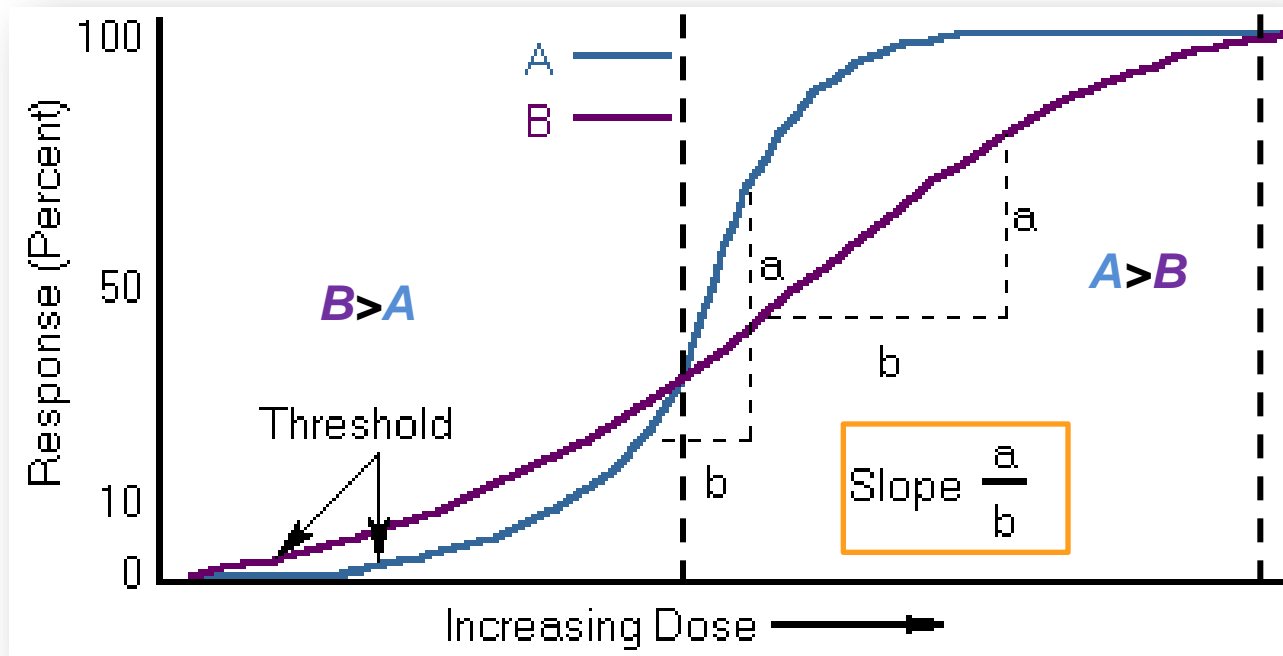
The dose below which there is no effect (dependent on the endpoint).



Between dose and response: a linear relationship or with a threshold...  
Dose vs response curves can take many different shapes.  
Different relative potency of two compounds defined by different shapes and slopes.

# Dose response curve: shape and slope

Slope= steep curve → small dose variations induce great differences in the effects.



Threshold dose = specific for endpoint and substance.

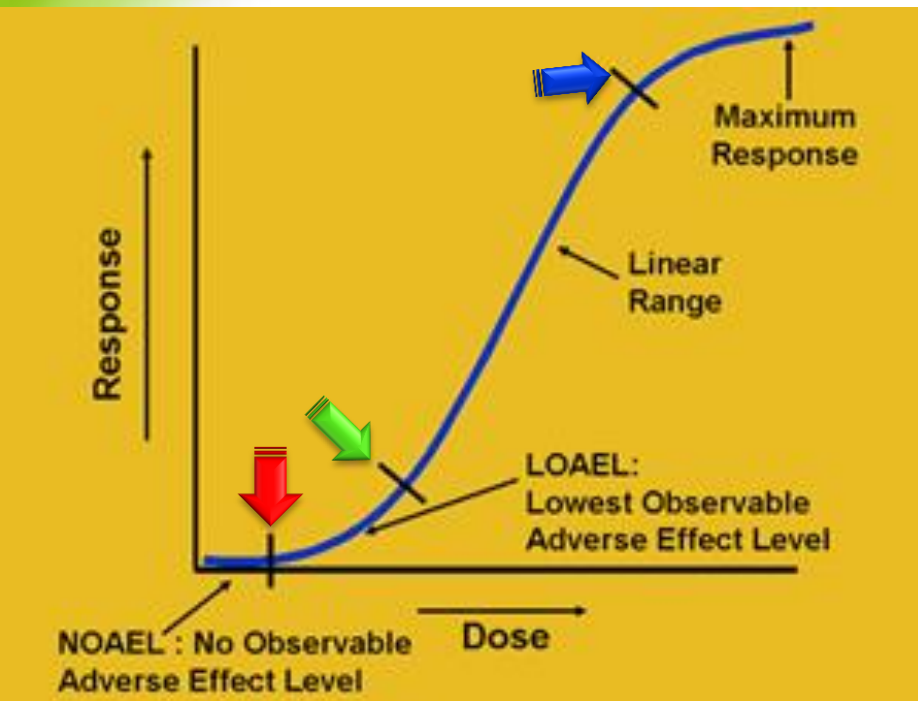
# Threshold dose and toxic effects

- A *systemic effect* is defined as an effect that is normally observed distant from the site of first contact, i.e., after having passed through a physiological barrier (mucous membrane of the gastrointestinal tract or of the respiratory tract, or the skin) and becomes systemically available.
- A *local effect* is an effect that is observed at the site of first contact, caused irrespective of whether a substance is systemically available.



All available information regarding systemic toxicity and local effects needs to be evaluated and, where possible, dose descriptors **N(L)OAE**, **Benchmark Dose (BMD)**, etc. need to be established.

# Dose descriptors/Reference Point (RP)



## NOAEL:

(No Observed Adverse Effect Level):

The highest exposure level at which there are no biologically significant increases in the frequency or severity of adverse effect between the exposed population and the control.

## LOAEL:

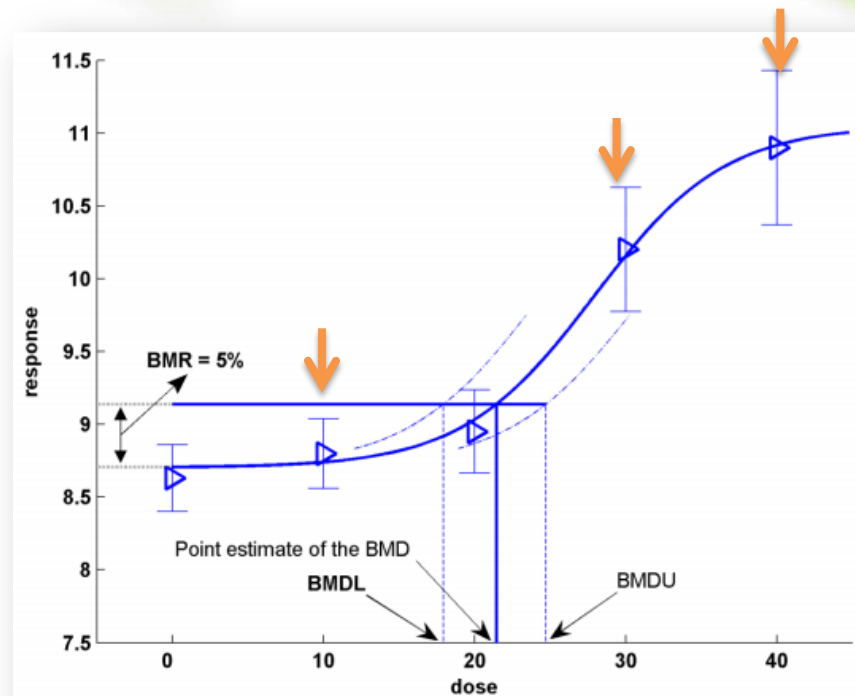
(Lowest Observed Adverse Effect Level)

The lowest exposure level at which there are biologically significant increases in frequency or severity of adverse effects between the exposed population and its appropriate control group.

# Dose descriptors/ Reference Point (RP)

**Benchmark Dose (BMD)**: is a dose level, derived from the estimated dose-response curve, associated with a specified change in response. BMDL is the BMD's lower confidence bound, and this value is normally used as the RP.

- A more **quantitative** alternative to the first step in the dose-response than the current NOAEL/LOAEL process.
- It is likely that there will continue to be endpoints that are not opened to modeling and for which a NOAEL/LOAEL approach must be used.
- In some cases a combination of BMDs and NOAELs



**Triangles**: observed mean responses +confidence intervals; **BMD** (point estimate); **Dashed curves**: upper and lower 95% confidence limits



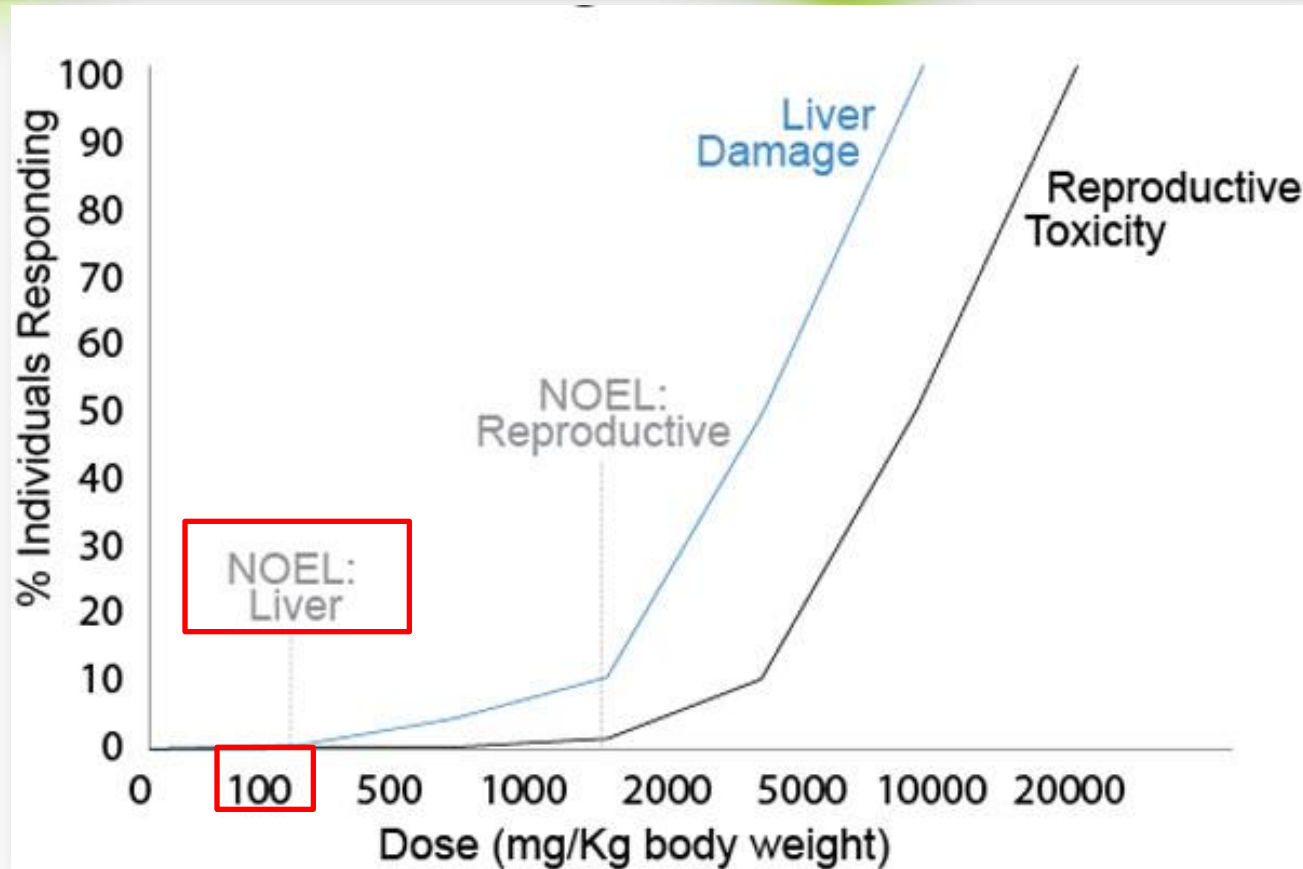
# Which is the critical effect?

## What questions are addressed within the test evaluation?

- Is it the one present at the lowest dose? n°/sex of animals?
- Is it the most relevant effect from a toxicological point of view (e.g. hair loss vs early marker of hepatic damage)
- Is the difference an effect of treatment?
- Dose –response? Due to outlier? Within historical control range? Biological plausibility? Consistency of the overall picture?
- Presence of related effects at higher doses or in the following times (indication of early effects) : alteration of biochemical parameters followed by histopathological changes (i.e. increase in hepatic transaminases vs hepatocellular hypertrophy or necrosis)
- Can the treatment related effect be defined as an adverse effect?

# Which is the critical effect?

For each endpoint → identification of a key study (data on ≠ species, ≠ exposure route, ≠ results from valid studies)



Preventing the critical effect we can avoid the other effects and thus we can get a total prevention for human health



# EXPOSURE ASSESSMENT

### III. Exposure assessment



Which is the level of exposure?

External 

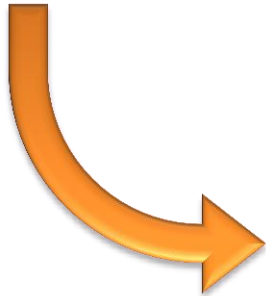
- At which concentration the chemical is present in different matrices or environment (diet/air/water/working place/consumer products)?
- Which is the preferred route(s) of exposure?
- Which the pattern of exposure? Working hours, environmentally, dietary.
- Which is the toxicologically relevant species? (e.g., parental or metabolite or degradation products).

# III. Exposure assessment

Internal



At what concentration the toxicologically relevant compound is present in different organ/tissue and in the target site?



## ADME

- Biomarkers of exposure
- Biomonitoring studies and “body-burden” measurements

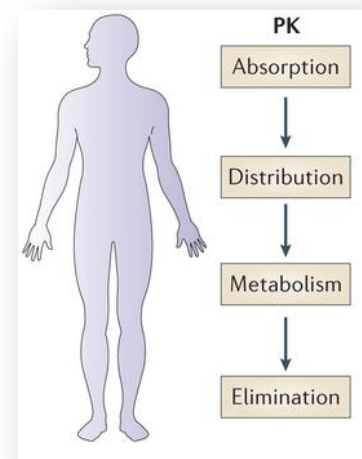
The most difficult dose to quantify: **target organ or biologically effective dose** (e.g. the dose that actually reaches the kidney).

Generally it cannot be directly measured. It is usually calculated on the basis of information gathered from TK studies.



# ADME studies: Parameters to be derived

- % **Oral Absorption**;
- % **Percutaneous Absorption** (if not available default values derived on the basis of MW and LogP);
- % **Inhalation Absorption** (if not available: route/route extrapolation);
- **Bioaccumulation Potential**

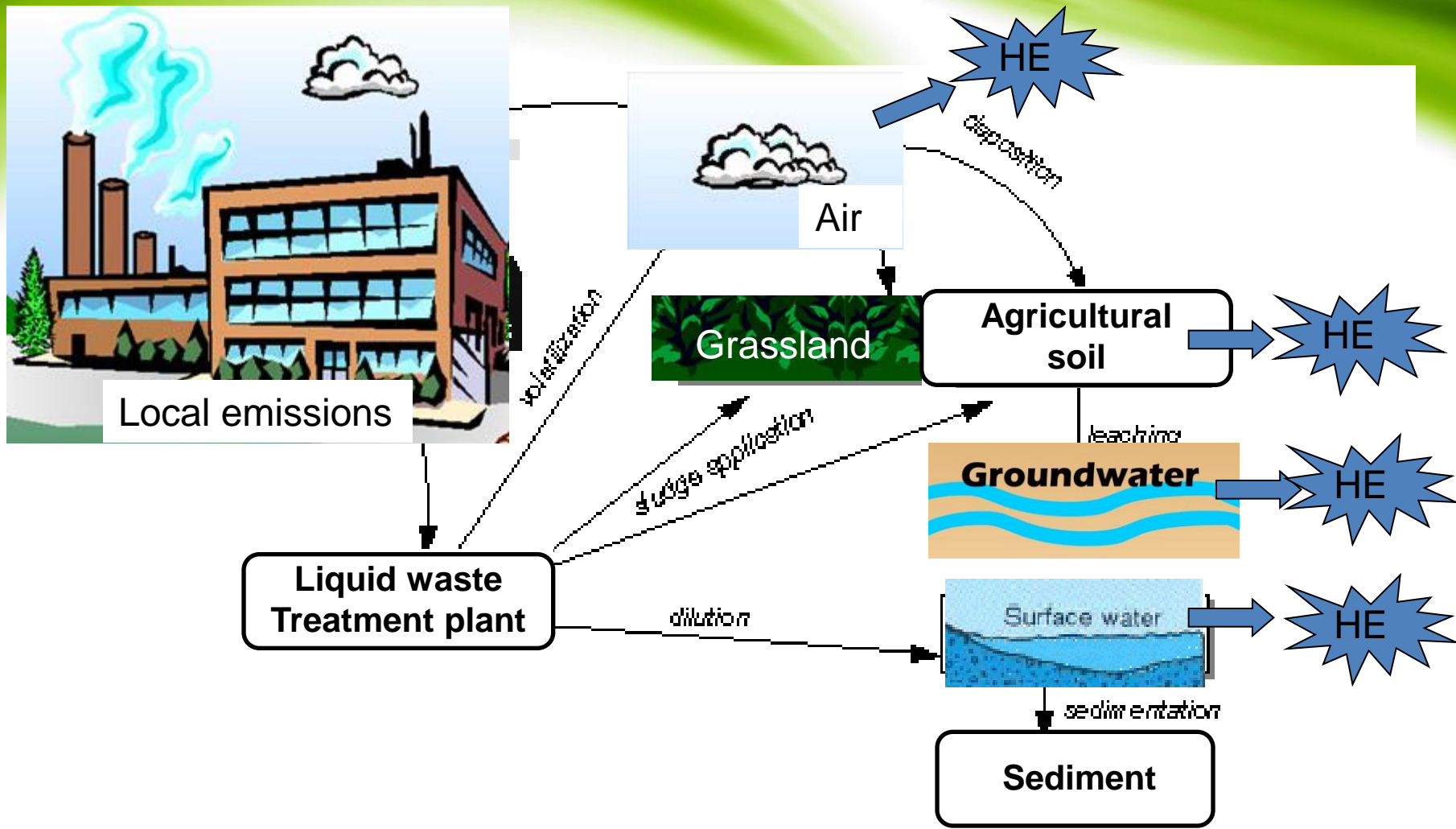


# Exposure pattern

Consumers; Workers, Operators, Bystanders;  
Indirect exposure

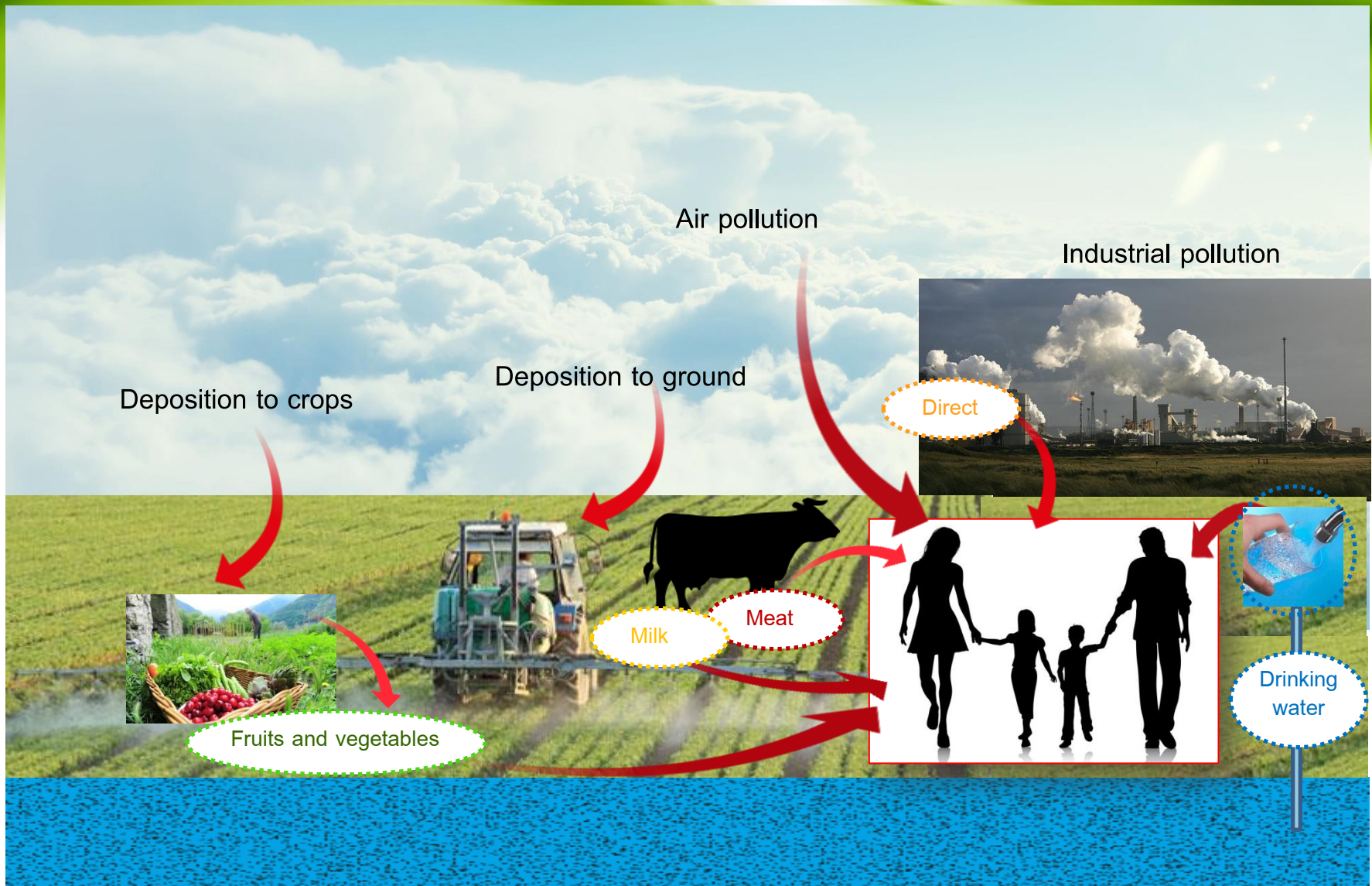


# Source of Exposure





# Pathways of exposure



# Exposure estimation

Xenobiotic exposure in humans can be measured in two ways, either through direct monitoring by measuring biomarkers from individuals or developing models to assess exposure.

- **Measuring Biomarkers**

Biomarkers, or biological markers, are chemicals/metabolites that can be measured in body fluid, such as urine, blood, saliva, and other body fluids.

- **Modeling**

It can be an alternative to estimate exposure to the direct measurements of biomarkers and less time consuming and expensive. It is basically a mathematical equation that inputs known variables to assess exposure levels.

The four main components of a model are environment, agent, host, and time.



# Exposure estimation: models

Occupational  
exposure model

Consumer  
exposure model

Oral



Dermal



Inhalatory

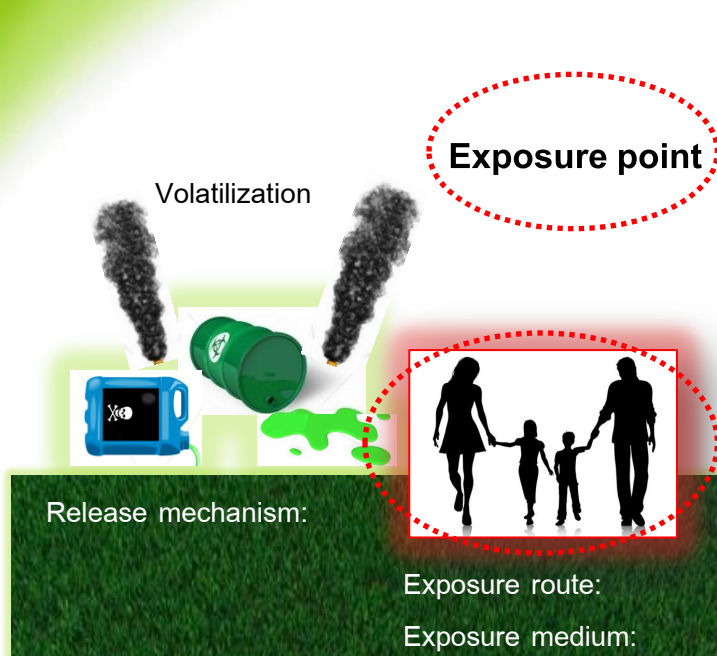


Four components that may be used in the model equation:

1. **Environment:** partition gradient, physical area;
2. **Agent:** chemical properties (e.g.vapor pressure, physical state, concentration);
3. **Host:** health, age, exposure pathway;
4. **Time:** length of exposure.

# Exposure estimation

## Scenario of exposure



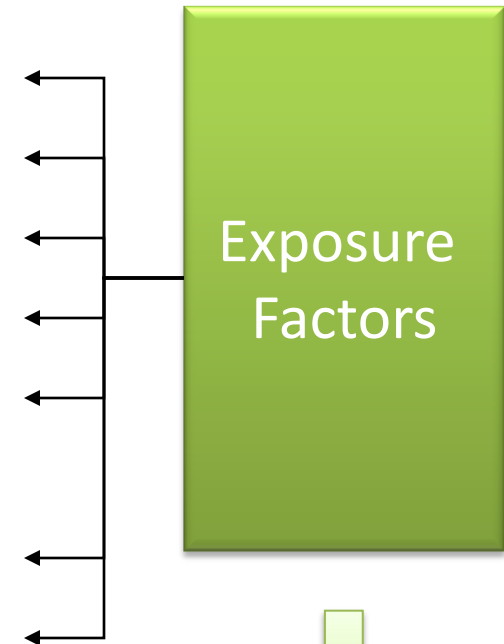
Description/  
Characterization

## Model



Formula conversion

## Model variables



Data to be  
insert into the  
model

# Exposure estimation: data quality

Body weight and height

Substance/formulation

physiologic data

use descriptions

dust and soil uptake

food consumption

time patterns

house and room characteristics

room ventilation

dermal uptake data

hand to mouth

emission rates

migration of substances in material



Level  
of  
quality



# RISK CHARACTERIZATION

## IV. Risk characterization



Which is the probability to have an effect and at what extent in the exposed population?

Having conducted the hazard assessment for all relevant human health endpoints and populations and the exposure estimation; a **quantitative risk characterization** is carried out.



Dose-response relationship data are compared with information of the **extent of exposure** to estimate the **probability to observe the toxic effect** within the population.

## IV. Risk characterization: Reference values (RfV)

Reference values are established for a given **critical effect**, and are specific to a substance, a duration of exposure (acute, subchronic /chronic) , a route of exposure (oral, inhalation, etc.).



Exposures above the RfV are not necessarily dangerous because a large **margin of safety** is allowed in their calculation but every effort should be made to keep the level of exposure below these values.



The **RfV** is an oral or dermal dose derived from:

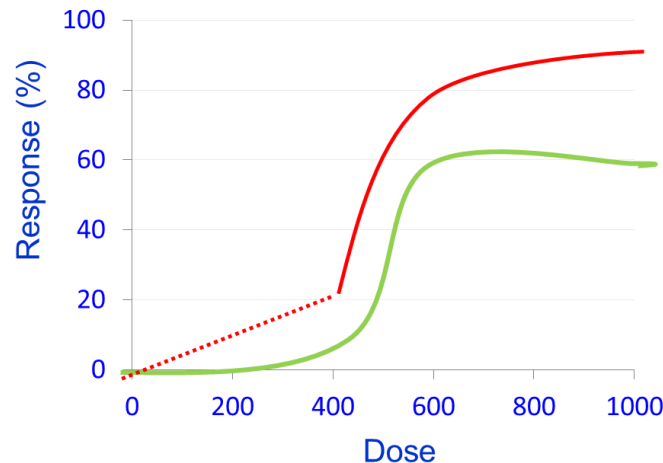
- the NOAEL, LOAEL or BMD
- by application of generally order-of-magnitude assessment/uncertainty/safety factors (Afs/Ufs/Sfs).



# Mode of Action (MoA)

Before deriving RfV:

- ? the substance exerts its effects by a **non-threshold MoA** (non-threshold mutagens or carcinogens);
- ? it is possible to derive a threshold



# Mode of Action (MoA) 1

There is a generally assumption that there is no threshold for safe exposure to substances which may cause cancer by mutation on DNA. If the substance exerts its effects entirely or partly by a **non-threshold MoA** (e.g. for mutagenicity, carcinogenicity) a RfV cannot be derived and for these effects semi-quantitative approach has to be followed where relevant or a qualitative approach for risk characterization:

e.g. Margin Of Exposure (MoE): to consider possible safety concerns arising from the presence in food of substances which are both genotoxic and carcinogenic.

It is a ratio of two factors: the dose at which a small but measurable adverse effect is first observed and the level of exposure to the substance considered.

# Non-threshold MoA → MoE approach

- Many substances are considered non-threshold compounds. The advice given by the risk assessor to the risk manager was to reduce exposure to a level that is as low as reasonably achievable. To overcome this, EFSA proposed the **margin of exposure (MoE)** approach.
- The **MoE approach** is not confined to genotoxic/carcinogenic compounds, but when the data are insufficient or inappropriate to establish a RfV.
- It uses a reference point (ex.  $BMDL_{10}$ : benchmark dose lower confidence limit 10% which is an estimate of the lowest dose which is 95% certain to cause no more than a 10% cancer incidence), taken from an animal study (but also from human data) corresponding to a dose that causes a low but measurable tumourigenic response. **This reference point is then compared with exposure estimates in humans.**
- EFSA considers that a **MoE of 10,000 or higher** (on  $BMDL_{10}$ ), would be of low concern for human health, to be considered as a low priority for risk management action.  $10,000 = 100$  (inter+intraspecies)  $\times 100$  (uncertainties in the nature of the carcinogenic process and the reference point).

## Mode of Action (MoA) 2

If the substance exerts its effects by a **threshold MoA**, the derivation of RfV on the basis of the Reference Point is required. RfV must be derived for the most critical effect(s):

it is often wise to focus on the most sensitive population; regulatory efforts are generally made to keep exposures below the population threshold, which is defined as the lowest of the thresholds (dose below which no adverse effect is expected to occur), of the individuals within a population.

## Reference values derivation/ Reference Point (RP) selection:

- Usually the study in the most sensitive and relevant species resulting in the most relevant lowest RP (e.g.: NOAEL(C)s, LOAEL(C)s, BMDs) will be selected for establishing the relevant RfV;
- A comparison of the relevant RP for RfV derivation for different time-frames provides useful information on the influence of exposure duration on the severity and spectrum of toxicity. An assessment of the entire data package is of high scientific value, as it helps in elucidating time-dependency of toxicity.

General rule: if several relevant NOAELs (or others) are available the one that would result in the lowest RfV for a given time-frame should be chosen. However the lowest RP may not always provide the lowest RfV as it depends on the Assessment Factors that will be used for its derivation.

Expert judgement

## Reference values derivation/ **Uncertainty factors**

- Uncertainty factors (UFs) (also called assessment factors, safety factors, adjustment factors or extrapolation factors) are used to derive health-based guidance values (RfV) by extrapolating from experimental animal data to humans;
- The setting of the overall UF is a critical step that involve a high level of expertise (with clear, scientifically based justification);
- UFs are intended to cover the **variability** and **uncertainty** arising out of such an extrapolation.



## Variability factors:



1. Exposure (duration, dose, route): which are the groups with higher levels of exposure?

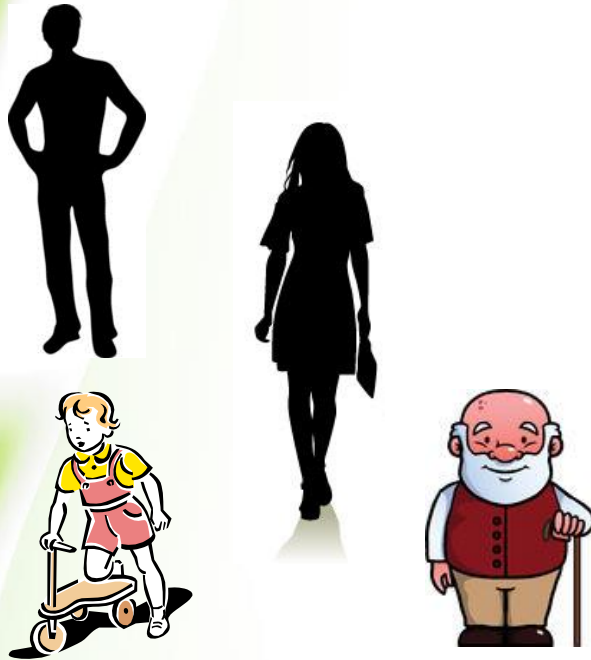


2. Susceptibility (age, patho/physiological status, genetic and/or acquired factors ): which are the more susceptible (vulnerable) groups, at the same levels of exposure?



**When the more vulnerable groups are protected, all the population is protected**

# Variability :



**Intraspecies differences** due to:  
age, gender, pathological status  
(liver/kidney/lung diseases),  
altered metabolic status due to  
exposure to contaminants  
(induction/inhibition of enzymes)

**Interindividual differences** in  
susceptibility due to  
genetic make up  
(enzyme polymorphism)



## Uncertainty:

- Quality of the available studies and of the experimental results: NOAEL vs LOAEL; relevance; data base consistency;
- Adequacy of the experimental model: relevant animal specie, study duration;
- Extrapolation of animal data → human  
high experimental doses → low actual exposure dose

## Overall UFs

- In the absence of chemical-specific data on kinetics and/or dynamics, it is recommended to use overall default UFs;
- If available and relevant, chemical-specific data on kinetics and/or dynamics and in general experimental results **should be used**.

# 1. Intra/inter-species extrapolation

Default uncertainty factor of **100**:  
10 for inter-species variability x  
10 for intra-human variability

$UF_{\text{inter-species}} = \text{TK:4} \times \text{TD:2.5}$  (WHO approach)

- for inter-species variability in toxicokinetics: 4.0
- for inter-species variability in toxicodynamics: 2.5
- for intra-human variability in toxicokinetics: 3.16
- for intra-human variability in toxicodynamics: 3.16

(EFSA approach)

*EFSA Journal 2012;10(3):2579*

An allometric scaling (to correct differences in the basal energetic metabolism: it changes among species) <sup>62</sup>



## 2. Other UFs

- **Deficiencies in the data available for the assessment:**

When additional data cannot be obtained or requested, the use of an additional UF to take account of the deficiency of a database should be considered on a case-by-case basis and justified. It is not possible to propose a default value for this UF, as it will be directly dependent on the dataset available.

- ✓ **Extrapolation for duration of exposure**

Extrapolation from subchronic to chronic study duration in rodents: EFSA recommends the use of an UF of 2.

- ✓ **Accounting for the absence of a NOAEL**

The LOAEL approach might be used and an additional UF will be needed (to be determined on a case-by-case basis).

- **Severity and nature of the observed effect:**

the need for an extra UF to allow for the severity of an effect is exceptional(case-by-case basis).

Ex. An irreversible and particularly severe toxicological effect: this should be a trigger to consider the finding in more detail before choosing an appropriate UF.



# Overall UFs

| Basis for AF                  | Factor | Source of uncertainty   |
|-------------------------------|--------|---|
| Human (Intraspecies)          | 1-10   | variability in human populations, (variation in sensitivity between members of the same species)  |
| Animal (Interspecies)         | 1-10   | differences in responses between animals and humans   |
| Subchronic to Chronic         | 1-10   | when chronic data are unavailable and a 90- day study is used   |
| LOAEL to NOAEL                | 1-10   | when LOAEL instead of NOAEL is used   |
| Data Gaps                     | 3-10   | applied for "Incomplete" data bases (i.e., adequacy of studies or database, missing studies)  |
| Nature and severity of effect | 1-10   |   |
| Modifying Factor              |        | usually applied for differences in absorption rates, tolerance to a chemical, lack of sensitive endpoint, or other toxicokinetics/dynamic parameters. The default value is 1. |

# To summarize:

## IDENTIFICATION OF CRITICAL EFFECTS:

Systemic and route specific effect; most relevant species and study; dose-response

THRESHOLD IDENTIFIED



DOSE RESPONSE ASSESSMENT



Identification of a Reference Point  
Select relevant NOAEL, LOAEL



MODIFICATION OF  
Reference Point  
Determination of absorption,  
bioavailability



APPROPRIATE UFs  
Chemical specific;  
Default values+additional UFs



REFERENCE VALUES

NO THRESHOLD IDENTIFIED



MoE approach

# Different Reference Values: some examples

- “The **ADI (Acceptable Daily Intake)** of a chemical is the estimate of the amount of a substance in food or drinking water, expressed on a bw basis, than can be ingested daily over a lifetime without appreciable health risks to the consumer on the basis of all known facts at the time of the evaluation” (WHO, 1997). Established for food additives and pesticide residues.
- **Tolerable Daily Intake (TDI)** is an estimate of the daily intake of a chemical contaminant which can occur over a lifetime without appreciable health risk.

Guidance Values (GV ): Maximum acceptable concentration in each exposure source (e.g.: drinking water, fish... ):


$$(TDI \times bw \times AF)/C$$

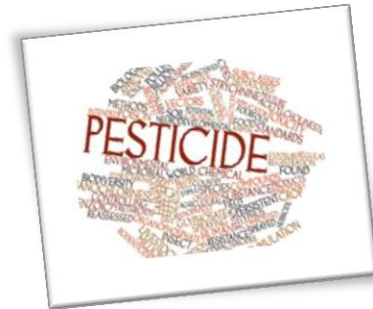
AF= allocation factor: fraction of the TDI allocated to each source;

C= daily consumption

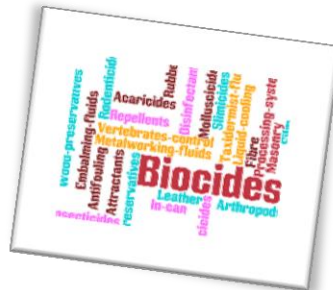
- **DNELs (Derived No-Effect Levels)** for threshold effects
- **AEL Accepted exposure level**

## Different Reference Values: some examples

- **Acceptable Operator Exposure Level (AOEL)** is the maximum amount of a.s., expressed on a b.w. basis, to which an operator may be exposed without any adverse health effects.
  - For highly acutely toxic compounds the reference value is the **Acute Reference Dose (ARfD)**.  
Daily consumption for a limited period of time.
- 



- Acute AEL



- Sub-chronic DNEL



# Derivation of ADI as an example

It is a health-based exposure limit to be used for comparison with estimated/measured exposure levels for assessing the risk also of the sensitive groups of population arising from the application of pesticides.

Normally based on Reference points (i.e. N(L)OAEs) that are obtained in long-term animal toxicity studies.



$$\text{ADI} = \frac{\text{N(L)OAE}}{\text{UF}} \quad (\text{inter} \times \text{intra} \times \text{expdur} \times \text{expert} \times \text{dose-resp})$$



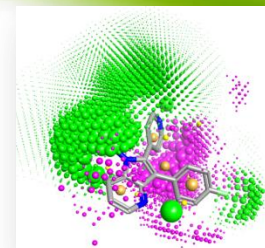
# Risk assessment: a nonstop evolution process



- To evaluate current approaches and alternative methodologies to the derivation of RfV. The basic considerations apply to a wide number of areas of chemical risk assessment. Evaluations have to be reviewed to determine studies and endpoints utilized to derive RfV, UFs applied and any aspects routinely debated during peer review.
- Alternative approaches to the NOAEL have to be evaluated. Particular consideration was given to the Benchmark Dose (BMD) approach.
- Alternative approaches to the use of the default 100 fold UFs to address uncertainties in extrapolating between animal data and human exposures.

# What are 3Rs alternatives?

- Efficient use of existing information;
- (Q) Structure-activity relationships, mathematical models, computer simulations;
- Grouping of chemicals, Read Across;
- *In vitro* methods:
  - isolated organs
  - tissue slices
  - tissue cultures
  - cell cultures
  - subcellular fractions;
- Lower animals; animal welfare



|                     |                                      |
|---------------------|--------------------------------------|
| <b>R</b> eduction   | - To minimize number of animals used |
| <b>R</b> eplacement | - To avoid the use of living animals |
| <b>R</b> efinement  | - To minimize suffering and distress |

# New challenges in risk assessment

## State of the art

Archives of Toxicology  
May 2011, Volume 85, Issue 5, pp 367-485

### Alternative (non-animal) methods for cosmetics testing: current status and future prospects—2010

Sarah Adler, David Basketter, Stuart Creton, Olavi Pelkonen, Jan van Benthem, Valérie Zuang, Klaus Ejner Andersen, Alexandre Angers-Loustau, Aynur Aptula, Anna Bal-Price, ... [show all 55](#)



### JRC SCIENCE AND POLICY REPORTS Alternative methods for regulatory toxicology – a state-of-the-art review

A. Worth, J. Barroso, S. Bremer, J. Burton, S. Casati, S. Coecke, R. Corvi, B. Desprez, C. Dumont, V. Goullarmou, M. Goumenou, R. Gräpel, C. Griesinger, M. Halder, A. Januschek, A. Kienzl, F. Madia, S. Munn, M. Nepelska, A. Paini, A. Price, P. Prieto, A. Rolaki, M. Schäffer, J. Triebel, M. Whelan, C. Wittwehr, V. Zuang



### SCIENTIFIC REPORT OF EFSA Modern methodologies and tools for human hazard assessment of chemicals<sup>1</sup> European Food Safety Authority<sup>2,3</sup> European Food Safety Authority (EFSA), Parma, Italy

EFSA Journal 2014;12(4):3638

2014

# New challenges in risk assessment

What are scientists doing to advance safety assessment of chemicals without relying on animal testing?



*Practical guide 10:*  
**How to avoid unnecessary  
testing on animals**



Scientific Committee on Health and Environmental Risks  
SCHER

Scientific Committee on Emerging and Newly Identified Health Risks  
SCENIHR

Scientific Committee on Consumer Safety  
SCCS

*Addressing the New Challenges for Risk Assessment*



# New challenges in risk assessment

☺ Exposure-based approaches such as TTC  
(Threshold of Toxicological Concern) for prioritization:

*“Toxicity testing in the 21<sup>st</sup> century: a vision and a strategy”;*



The **Threshold of Toxicological Concern (TTC)** is a principle which refers to the possibility of establishing a human exposure threshold value, below which there is no appreciable risk to human health.

- For each structural classes of chemicals by considering extensive databases of toxicity data (by the oral route) generated in the past.
- Requires reliable exposure assessment and confidence in data evaluation.

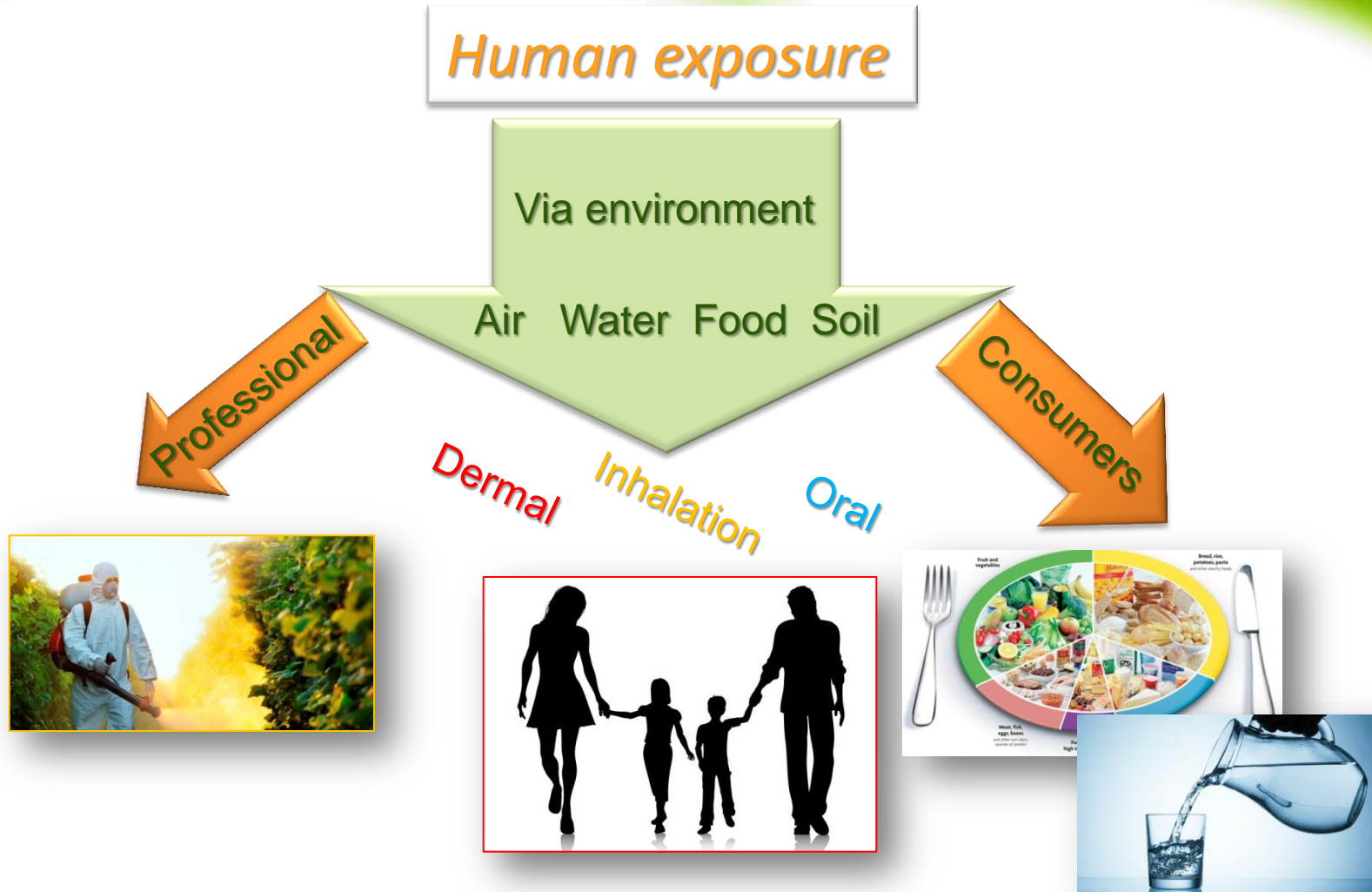


# CASE STUDIES



# Exposure pattern

Consumers; Workers, Operators, Bystanders;  
Indirect exposure



# 1. Pesticides: dietary exposure assessment

Dietary exposure to pesticide residues =  
Food consumption  $\times$  Pesticide residues levels in food



Human Health Risk Assessment



Dietary  
exposure

Reference values  
(ADI; ARfD)

# 1. Pesticides: dietary exposure assessment

- **Acute exposures:** calculated over a period of one day;
- **Chronic exposure:** assessed as the average daily exposure of an individual over their lifetime;
- It considers the whole of the EU or national population. EU Regulation requires particular attention to protection of **vulnerable groups** including pregnant women/unborn children, infants and children → by conducting specific exposure assessments for vulnerable groups, or by assessing the overall population (separated results for vulnerable groups);
- Consideration of **all plant and animal commodities** in the form they are consumed (raw and/or processed) when they are expected to contain residues of the pesticide in question, and all foods that contain those commodities;
- EU regulation requires that account should be taken **of known cumulative and synergistic effects** where scientific methods to assess such effects are available;



# 1. Pesticides: dietary exposure assessment

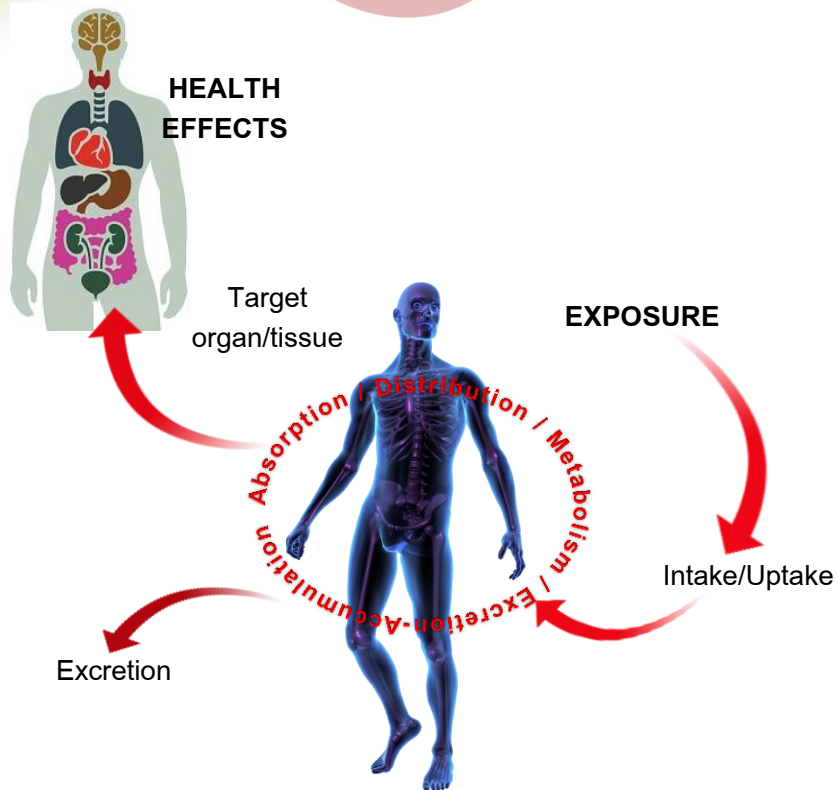
- Available residue data are combined with dietary information to estimate potential residue intake by consumers. The consumer is considered to be adequately protected when **estimated dietary intake** of pesticide residues **does not exceed** the acceptable daily intake (ADI) or the acute reference dose (ARfD).
- The health impact of pesticides in food is estimated by comparing **dietary exposure to toxicological levels of concern (RfVs)**.  
Exposure assessments combine data on concentrations of a pesticides present in food with the quantity of those foods consumed.
- Theoretical Maximum Daily Intake (TMDI); International Estimated Daily Intakes (IEDIs); National Estimated Daily Intakes (NEDIs).
- The risk assessment methodology developed for this specific risk assessment task is based on internationally recognized methodologies → [EFSA model for chronic and acute risk assessment](#)



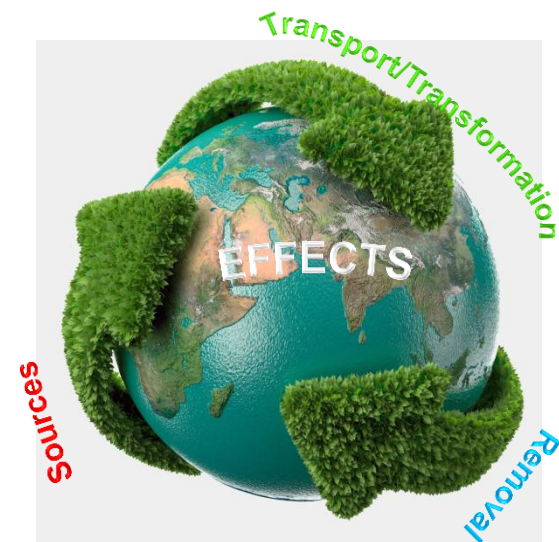


# Human vs Environmental Health

Human  
Health Risk  
Assessment



ENVIRONMENTAL health



# Environmental Health Risk Assessment

## Environment and health :

- **Key points:** some characteristics of environmental pollution and the variety and heterogeneity of the possible exposure scenarios for the people living in polluted areas.
- The process for estimating exposures to environmental pollutants is one of the most important sources of uncertainty, limiting the ability to quantify the role of the environment as a determinant of human health.
- **Analytical process:** the characterization of environmental exposure must follow an integrated multi-disciplinary approach: characterization of the **sources of emission/release**; **distribution** of sources of contamination; **time-frame** of contamination;
- **Transport , environmental fate and bioavailability of contaminants:** Physico-chemical properties play a key role in the analysis of the environmental fate of pollutants and transport.



The aim is to describe: What compartments of the ecosystem are most affected; the factors that influence the environmental concentrations of a contaminant , its bioavailability and its potential to bioaccumulate and to produce break down products (secondary pollutants).



# Environmental fate and behaviour

Required data for the estimation of exposure, for example:

- **Source** : Levels of the substance in a product, its release, use.
- **Fate**: Identification of release , distribution , adsorption and desorption , degradation.
- **Contact**: behavior of people exposed , contact time , frequency , duration (e.g.: children crawling, playing on the ground, sucking toys ).
- **Routes** by which pollutants enter ecosystems (e.g. surface waters).



The **physico-chemical properties** of the pollutants and the characteristics of abiotic compartment determine how long and how the substance is or move from/to a given compartment:

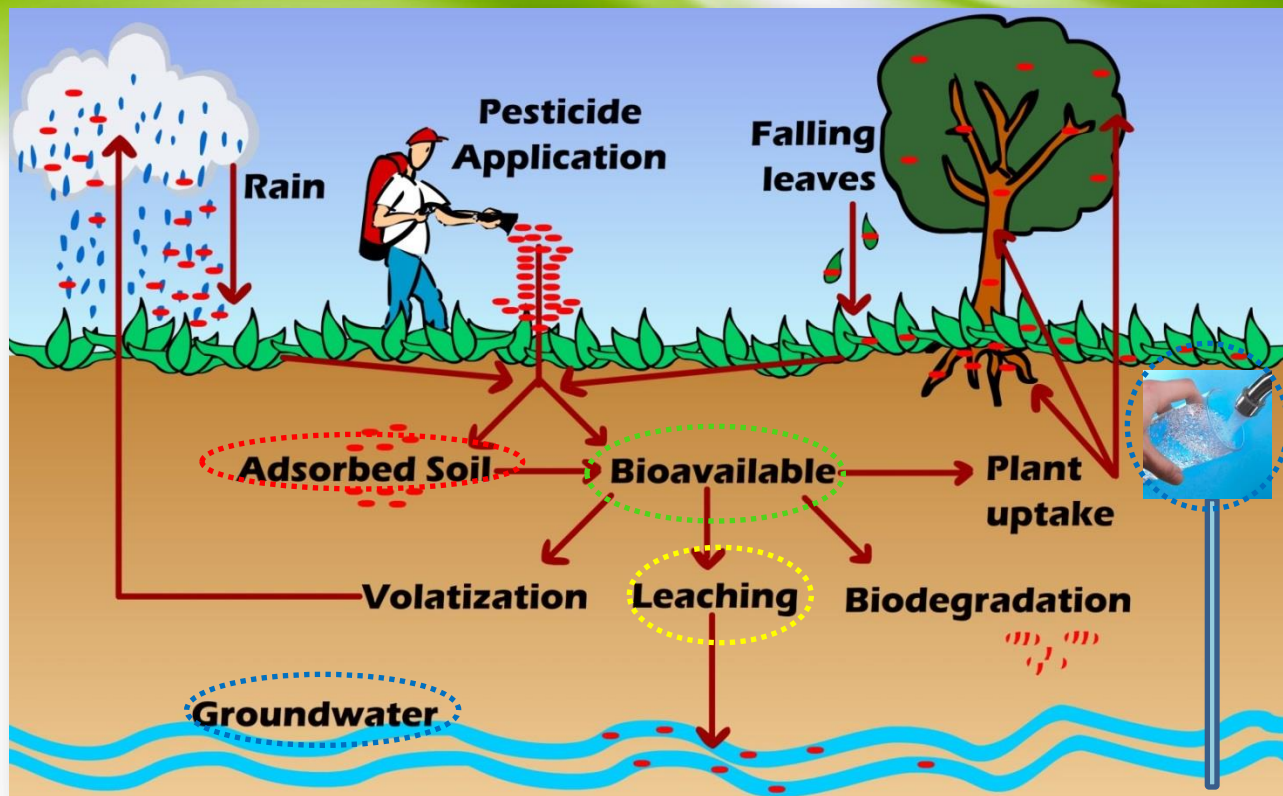
**Molecule properties:**

mass, charge, solubility, vapor pressure, partition coefficients;

**Properties of the compartments:**

pH, organic matter, soil quality, fine particulate , temperature, etc.

## 2. Pesticides: in the environment



*Persistence and Bioaccumulation of Persistent Organic Pollutants (POPs); <http://dx.doi.org/10.5772/56418>*

### Application Rate Approach: Plant Protection Products

- **Soil exposure** is directly dependent on application rate (as defined by the GAP, Good Agricultural Practices).
- Also determines **quantity available to be transported via other pathways**.
- **Run-off:** movement in water over a slopping surface.

## 2. Pesticides: in the environment

### Environmental fate data:

- Generally relates to the calculation of **predicted environmental concentrations (PECs)** in all relevant compartments. This can involve relatively simple calculations for **soil** or **water** via spray drift, or more complex modelling using agreed European tools such as the FOCUS models to simulate long term leaching behavior to **groundwater**.
- Soil **metabolites** which may leach into groundwater must also be assessed as for the active substance.



Metabolites which are found to occur at **significant levels ( $>0.1 \mu\text{g/l}$ )** must also be considered for **toxicological relevance** in a stepwise approach according to the approaches in the

*“Relevance of metabolites in groundwater, SANCO/221/2000 – rev.10, 25/02/2003”*

## 2. Pesticides: relevant metabolite in groundwater

### Sequential assessment of the relevance of metabolites:

- **Step 1** (fate and behaviour expert): Exclusion of degradation products of no concern, like CO<sub>2</sub> or an inorganic compound, not containing a heavy metal;
- **Step 2** (fate and behaviour expert): Quantification of potential groundwater contamination. FOCUS groundwater models and scenarios have to be used with data on degradation and sorption of metabolite as input;

If a metabolite is predicted in groundwater at a **concentration higher than 0.1 µg/L**, it has to be identified as relevant or not by toxicologists.



- **Step 3** (human toxicology expert):  
*Hazard Assessment*: Identification of relevant metabolites.
  - Screening for biological activity,
  - Screening for genotoxicity,
  - Screening for toxicity.

If a compound is considered **relevant**, the threshold of 0.1 µg/L is of reference.  
If a compound is considered **not relevant** with respect to hazard assessment, the exposure assessment (Step 4) has to be evaluated:





## 2. Pesticides: relevant metabolite in groundwater

- **Step 4** (fate and behaviour/toxicology expert):  
*Exposure assessment-TTC approach.* For non-relevant metabolite a **TTC of 0.75 µg/L** can be considered for GW. If the **PEC<sub>gw</sub>** is < of this limit, no concern is expected. If not, the Step 5 has to be taken into account.
- **Step 5** (human toxicology expert): *Refined risk assessments for non-relevant metabolites.*

Non-relevant metabolites with **0.75 < PEC<sub>gw</sub> < 10 µg/L** require a **refined assessment** of their potential toxicological significance risk for consumers.

*Case-by-case basis:*



For example: for a metabolite found in animals the acceptable limit may be defined starting from the ADI values derived for the parent compound.

The MAC (Maximum Acceptable Concentration) can be evaluated as:

$$MAC = \frac{ADI \times bw}{C_w \times LF}$$

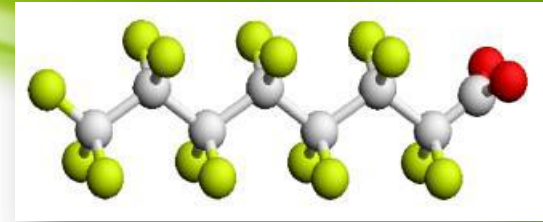
where:

ADI= Acceptable daily intake; *bw* = 60 kg for an adult; *C<sub>w</sub>* = daily water consumption (2L);  
LF = location factor of 10% (% ADI attributable to drinking water)

### 3. Perfluoroalkylated substances (PFAS)



#### Environment and health



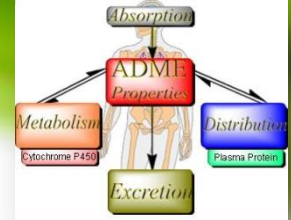
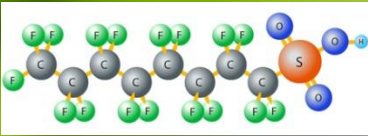
- **Perfluoroalkylated substances (PFAS)**: is the collective name for a vast group of fluorinated compounds, including **PFOA** is perfluorooctanoic acid and is sometimes called C8 or C-8, perfluorooctane sulfonate (**PFOS**) .
- **PFAS** have been widely used in **industrial and consumer applications** including stain- and water-resistant coatings for fabrics and carpets and clothing (GORE-TEX®), oil-resistant coatings for paper products approved for food contact, fire-fighting foams, floor polishes, and non-stick surfaces on cookware (Teflon®).
- However, consumer products made with fluoropolymers and fluorinated telomers, including Teflon® are not PFOA. Rather, some of them may contain trace amounts of PFOA and other related perfluorinated chemicals as impurities. Therefore, a number of different PFAS have been widely found in the environment





# 3. Perfluoroalkylated substances (PFAS)

## Physico-chemical and kinetic characteristics



- **vP, TvB, BB and T** (that is very persistent , very bioaccumulative and toxic). The bioaccumulation potential depends on the level of exposure and can hardly be traced to a single value of BAF ( bioaccumulation factor).
- Data on the proportion between body-burden and age are conflicting .
- The potential for **bioaccumulation of PFAS** is not related to the partition in lipid tissue and adipose tissue, but
  - ? **rapid oral absorption**
  - ? **relevant protein binding** in plasma , accumulation in the liver, kidney, absence of biotransformation processes
  - ? **slow elimination** with re-absorption at renal level



### 3. Perfluoroalkylated substances (PFAS)

Reference values for PFOS e PFOA according to different  
Regulatory agencies

|                 | <i><b>PFOA</b></i>  | <i><b>PFOS</b></i>   |
|-----------------|---------------------|----------------------|
| <i>EFSA</i>     | <i>1,5 µg/kg bw</i> | <i>0,15 µg/kg bw</i> |
| <i>U.S. EPA</i> | <i>0,4 µg/L</i>     | <i>0,2 µg/L</i>      |
| <i>UK</i>       | <i>3,0 µg/kg bw</i> | <i>0,3 µg/kg bw</i>  |

## PFOS TDI in food chain

EFSA Journal 2008; 653: 7-131

EFSA Journal 2012;10(6):2743

- From a subchronic study in *Cynomolgus* monkeys, the *Scientific Panel on Contaminants in the Food Chain (CONTAM)* identified **0.03 mg/kg b.w.** per day as the lowest no-observed-adverse-effect level (**NOAEL**) and considered this a suitable basis for deriving a Tolerable Daily Intake (TDI).  
**End-point:** Biochemical alterations (lipid metabolism, thyroid hormones);
- The CONTAM Panel established a **TDI** for PFOS of **150 ng/kg b.w. per day** by applying an overall uncertainty factor (**UF**) of **200** to the NOAEL.  
An UF of 100 was used for inter and intra-species differences and an additional UF of 2 to compensate for uncertainties in connection to the relatively short duration of the key study and the internal dose kinetics.

**TDI: 150 ng/kg bw/day**  
(0.03 mg/kg : 200)

## PFOA TDI in food chain

*EFSA Journal 2008; 653: 7-131*

*EFSA Journal 2012;10(6):2743*

- The CONTAM Panel noted that the 95% lower confidence limit of the benchmark dose for a 10% increase in effects on the liver (**BMDL<sub>10</sub>**) values from a number of studies in mice and male rats were in the region of **0.3 - 0.7 mg/kg b.w./day**.  
Therefore, the Panel concluded that the **lowest BMDL<sub>10</sub> of 0.3 mg/kg b.w./day** was an appropriate point of departure for deriving a TDI.
- The Panel established a **TDI for PFOA of 1.5 µg/kg b.w./day** by applying an overall **UF of 200** to the BMDL<sub>10</sub>.  
An UF of 100 was used for inter- and intra-species differences and an additional UF of 2 to compensate for uncertainties relating to the internal dose kinetics.

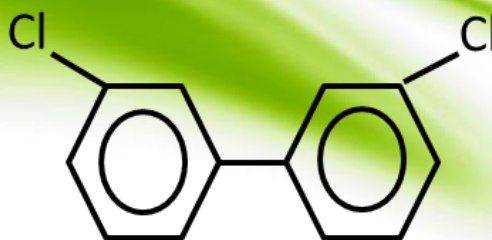
**TDI: 1,5 µg/kg bw/day**

(0,3 mg/kg : 200)



The CONTAM Panel noted that the indicative human average and high level dietary exposure for PFOA of 2 and 6 ng/kg b.w. per day, respectively, are well below the TDI of 1.5 µg/kg b.w. per day.

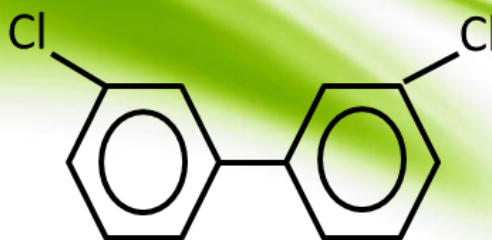
## 4. Persistent Organic Pollutants (POPs)



PCB

- Polychlorinated biphenyls (PCBs) are ubiquitous environmental pollutants widely used as industrial chemicals, particularly in the electrical industry and for the manufacture of paints, plastics and adhesives. They have been in use since the beginning of the 1930's;
- It is a mixture of compounds containing the biphenyl structure with varying numbers (1 to 10) and arrangements of chlorine atom attached. PCB'S are generally used as a mix of isomers;
- Due to PCBs' environmental toxicity and classification as a persistent organic pollutant (POP), PCB production was banned some years ago.

## 4. PCBs properties

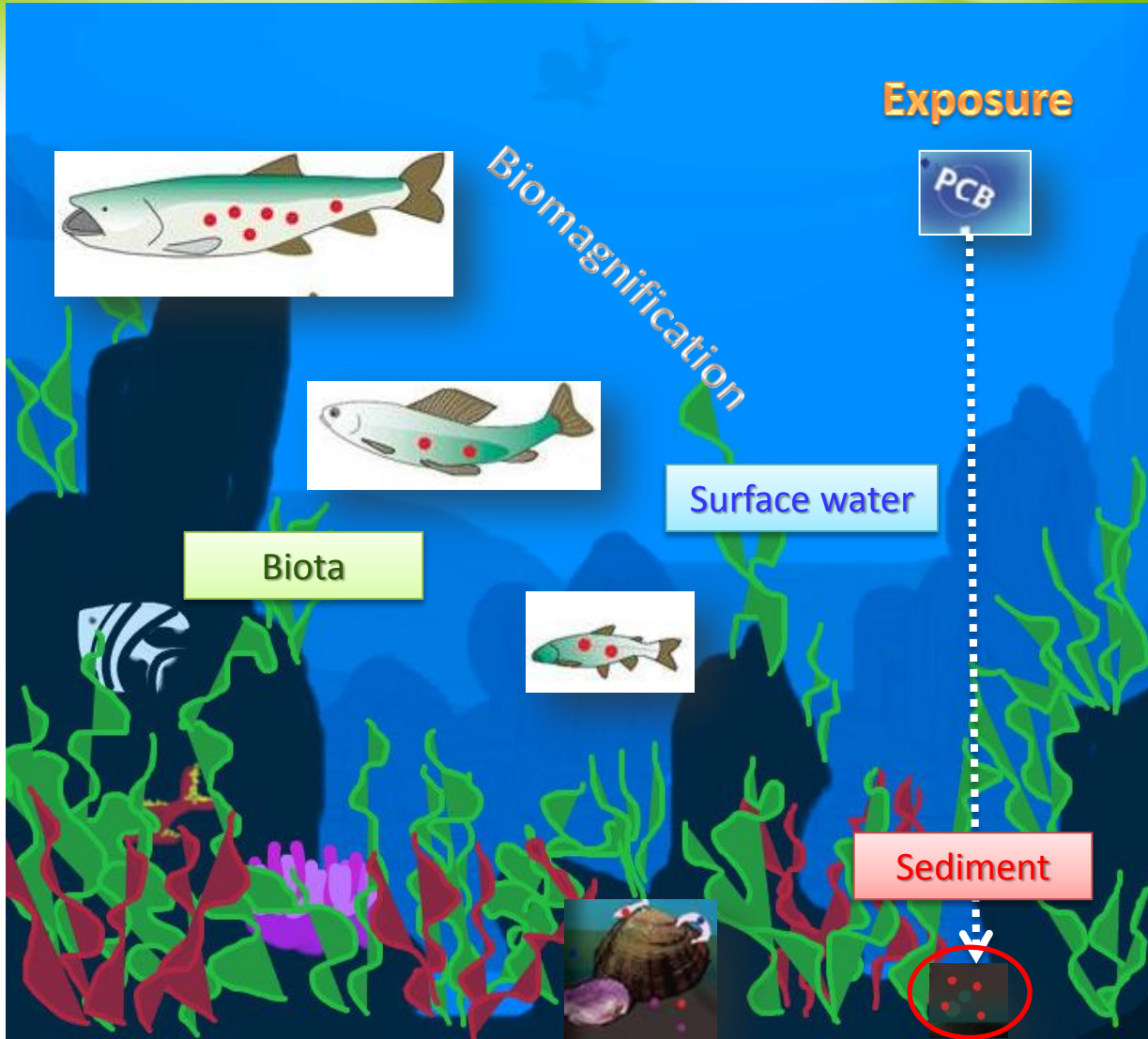


PCB

- **Chemical properties:**
  - I. lipophilic (easily mixed with oil and fat): high Log Kow (6.0-8.2);
  - II. very stable;
  - III. low aqueous solubility:  
48.6 (PCB 47), 5.97 (PCB 99), 1.28(PCB 153)  $\times 10^{-6}$  ;
  - IV. low vapor pressure;
- **Persistent** in the environment;
- **Bioaccumulation** & **bioconcentration** effects;
- **Accumulation in human fat tissue:**  
i.e. humans: 2300 ng/g, human breast milk: 1200 ng/g.



#### 4. PCBs: bioaccumulation and biomagnification



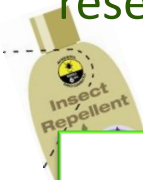
## 4. PCBs: contaminated site Risk Assessment

- PCBs can be carried in the atmosphere, changing phases depending on temperature and vapor pressure, to contaminate **surface water or soil**;
- **Living organisms** : important **source** of POPs, contaminating plants and small organisms that are then consumed by larger animals. Progression up the food chain, with each animal consuming greater quantities of contaminated species on the lower levels of the food chain, **magnifies the concentration of the contaminant consumed (biomagnification)**;
- **Routes of exposure**: dermal (skin absorbed), respiratory (inhaled), and gastrointestinal (ingested);
- **Pathways of exposure**: through soil ingestion (oral); through vegetable and food consumption (oral); through inhalation of indoor air (inhalation);
- **Assessing human exposure**: i. measurement of the actual body burden through **biomonitoring** (sampling and measuring body fluids or body tissue). For several reasons, however, these measurements often offer limited possibilities and are only used in higher tier Risk Assessments (i.e. importance of reliability of the data); ii. by using a so-called **exposure model**. These exposure models enable calculation of the rate contaminants enter the human body, blood stream, or target organs.

# Risk assessment: lessons learned



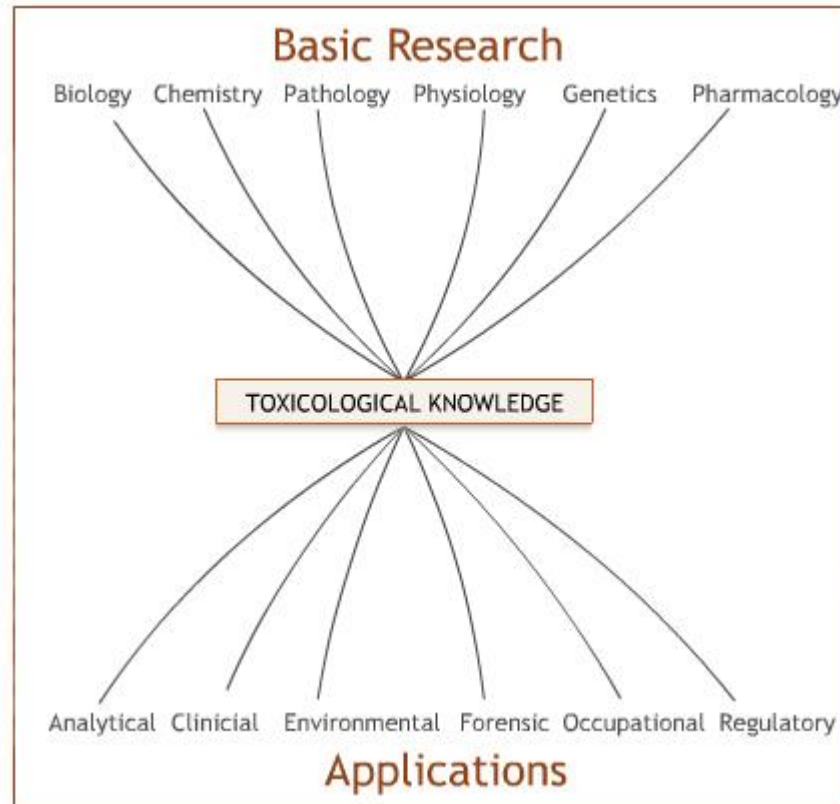
- **Consistent/transparent**: to ensure objectivity and rationality;
- **easily understood** by all the interested parties; flexible enough to deal with complex situations; reproducible;
- based on the **best scientific evidence** available at the time, **well-documented** and supported with references to the scientific literature/data and other sources, including **expert opinion**;
- regularly reviewed and **updated** when additional new information becomes available;
- complemented by decisions/actions based on available information; taking into account **uncertainties** (gaps in knowledge) and **assumptions** made, in order to evaluate the effect of these on the final risk estimate and priorities for future research.



**Fundamental principles of Risk Assessment apply to**  
Plant Protection Products, Biocides, Chemicals within REACH, Cosmetics,  
Pharmaceuticals, Environmental Contaminants...

# What to do now?

You have now been introduced to some of the basic principles of toxicology.





# Some useful web sites

**OECD:** <http://www.oecd.org>

*Test Guidelines Series on Testing and Assessment;*

**EFSA:** <http://www.efsa.europa.eu/en/panels/pesticides.htm>;

**EPA:** <http://www.epa.gov/>

**Echa:** <http://echa.europa.eu/home>

**European commission:**

[http://ec.europa.eu/sanco\\_pesticides/public/?event=homepage](http://ec.europa.eu/sanco_pesticides/public/?event=homepage);

**CircaBC:**

[https://circabc.europa.eu/faces/jsp/extension/wai/navigation/contain  
er.jsp](https://circabc.europa.eu/faces/jsp/extension/wai/navigation/contain<br/>er.jsp);

**ISS:** <http://www.iss.it/>

**JRC- EURL ECVAM:** [http://ihcp.jrc.ec.europa.eu/our\\_labs/eurl-ecvam](http://ihcp.jrc.ec.europa.eu/our_labs/eurl-ecvam)



***Thank you very much  
for Your Attention!!***

